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1 **Extended performance study of forward osmosis during wastewater**
2 **reclamation: Quantification of fouling-based concentration**
3 **polarization effects on the flux decline**

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27 **Abstract**

28 The long-term performance of forward osmosis during simulated wastewater reclamation was
29 investigated for 120 h operation with a focus upon the influence of flux on flux decline and the
30 synergistic effect of fouling on concentration polarization. Our comprehensive investigation
31 focused on different fluxes (25; 30; 34 LMH) for simulated wastewater containing either a high
32 protein or a low protein fraction. Compared to an initial flux of 25 LMH, operation at an initial
33 34 LMH favored the formation of a thicker and more compact cake layer which resulted in
34 significant increase in both cake structural parameter (four-fold) and cake layer enhanced
35 concentration polarization (ten-fold). After 40 h operation without physical cleaning the additional
36 effect of cake layer enhanced concentration polarization and fouling resistance consumed 25% of
37 the total driving force; the significant internal concentration polarization still had the greatest
38 impact. In contrast operation at the lower flux of 25 LMH generated less fouling with a lower cake
39 structural parameter (119 μm). The resultant flux decline was only 3% in contrast to the 15-18%
40 found for the higher flux of 34 LMH. For operation above an initial 30 LMH it was found that FO
41 fouling became irreversible if the wastewater contained a high protein fraction. Overall for a thin
42 film composite membrane and a wastewater with a foulant concentration of 160 mg/L an initial
43 flux of 25 LMH is the recommended threshold; this is 25% less than the critical value determined
44 in earlier short-term studies.

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46 **Keywords:** Forward Osmosis; Water reclamation; Long-term operation; Flux decline pattern.

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54 **1. Introduction**

55 To resolve the water scarcity issue, seawater desalination and wastewater reuse are two main ways
56 to supply fresh water. Due to the limitation imposed by thermodynamics, further significant
57 reductions in the energy consumption of seawater reverse osmosis (SWRO) will be modest. Thus
58 the range of water production cost can be expected to remain relatively high at 0.5-2 USD/m³ [1].
59 Consequently, either primary or secondary wastewater effluent could be feasible sources for water
60 recovery and reuse [2]. Water cost depends on the purpose of usage i.e. direct/indirect potable or
61 non-potable reuse, ranging from 0.4-1.26 USD/m³ [2]. Pressure membrane filtration processes are
62 generally applied for water reclamation. However, these processes entail a high energy demand
63 due to the use of a transmembrane pressure generated by the pump, especially for reverse osmosis
64 (RO). Additionally, when wastewater is directly entered into the RO process, the irreversible
65 fouling is anticipated to occur [3]. Forward osmosis (FO) membrane is emerging as a feasible
66 technology for water reuse due to its energy-efficient and apparently lower fouling potential [4].
67 For a discussion of the later, see [22]. Whilst several studies have focused on the stand-alone FO
68 for reclaiming primary/secondary wastewater effluent [5] [6] [7], even for high strength
69 wastewater [8] [9], the FO membrane can be integrated with other processes (RO, UF, MD) as a
70 means of osmotic dilution to attain simultaneous water reclamation and desalination [10] [11] [12]
71 [13].

72

73 Previous work has indicated the cellulose triacetate (CTA) FO membrane has a superior
74 performance being less susceptible to fouling and with higher flux recovery (60-98%) [14] [15]
75 compared to the TFC membrane (30-50%) [5] [16]. However, these works only investigated a
76 single foulant model such as sodium alginate (ALG), bovine serum albumin (BSA), or humic acid
77 (HA) [17] [18]. In practice, wastewater contains complex organic foulants i.e., protein,
78 carbohydrate, and natural organic matters, which leads to severe fouling and thus hinder
79 sustainable long-term operation. As previously reported, biopolymers are viable foulants inducing

80 irreversible fouling in the pressurized membrane-based water treatment process [19]. This has also
81 been reported in a past study on FO-RO hybrid process-based direct potable reuse [20]. Another
82 study that used secondary wastewater as feed in the long-term FO operation reported that there
83 was continual irreversible fouling [21]. The degree of FO fouling, cake layer compaction, and
84 fouling reversibility are strongly influenced by the initial water flux [16] [22] [23]. Other works
85 have indicated that FO fouling is very complicated because a greater amount of fouling does not
86 necessarily lead to a greater severity of fouling [24]. As a final comment it is noted that FO fouling
87 not only causes a decrease in water flux, change in product quality but also the membrane is
88 damaged by the consequential cleaning.

89
90 As the inherent drawback of the FO membranes are concentration polarizations (CPs) effect,
91 including dilutive internal concentration polarization (ICP) inside the porous membrane and
92 external concentration polarization (ECPs) on both feed and draw side [25], a study of the
93 synergistic effect of fouling on concentration polarization is important. The CPs cause a decrease
94 in effective driving force and this fact is attributed to both accumulation/concentration of solute in
95 the feed solution and the dilution of draw solution in the porous layer [26]. Past studies have
96 evaluated the effect of CPs on overall FO performance [26] [27] [28]. Whilst their findings
97 indicated CPs became more serious as DS concentration increased, these effects may be mitigated
98 to a certain extent with increasing cross-flow velocity (CFV) i.e., from 5.56 to 11.11 cm/s [27],
99 from 2.11 to 36.40 cm/s [28]. It has been found that fouling diminished by elevating CFV i.e.,
100 from 10.7 cm/s ($Re= 615$) to 32.1 cm/s ($Re= 1936$) [11] or from 6 to 24 cm/s [29] ; however, this
101 option will cause an increase in energy consumption [8] [30]. Although the effect of CPs has been
102 analyzed for the FO membrane process, e.g. [27] [28], fewer studies have investigated their
103 influence when fouling occurs. When a fouling cake layer is formed on the active layer surface,
104 this hinders back-diffusion of the solution and results in an increase in osmotic concentration at
105 the interface between cake layer and active layer (AL). For the RO process, this has been termed

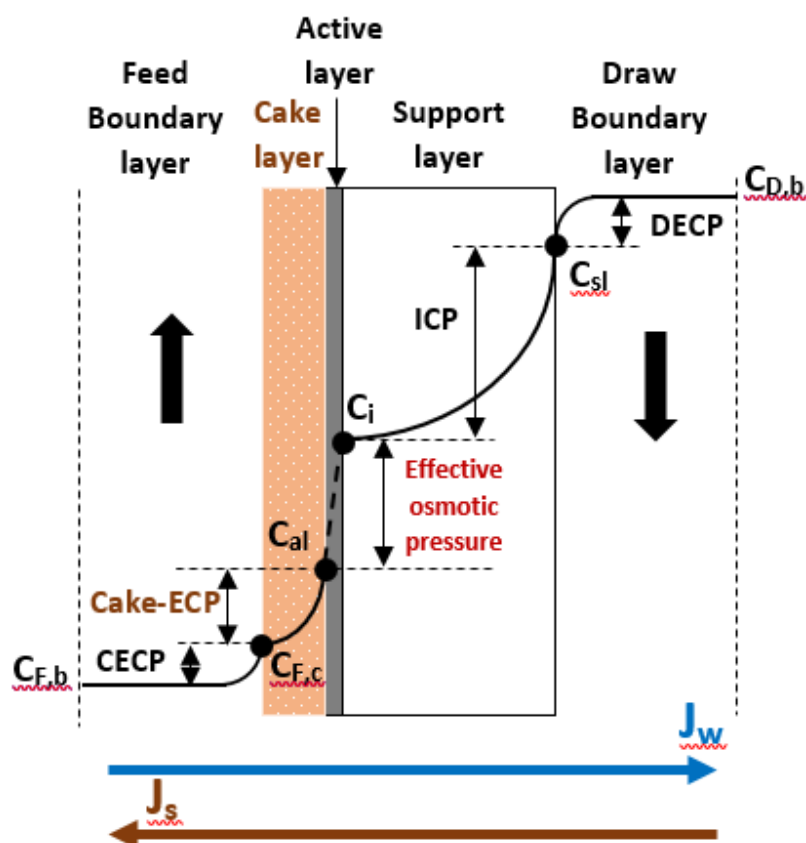
106 cake-enhanced concentration polarization (C-ECP) [31]. Now, with regard to the FO process, this
107 phenomenon can be labelled as cake-enhanced osmotic pressure (C-EOP) because the solute
108 accumulation induced by the effect of reverse solute flux (RSF) gives an elevated level of osmotic
109 pressure within the fouling layer. The C-ECP effect has been reported in several studies of FO
110 membrane fouling [29] [32]. However, it is worth noting that their works did not quantify its
111 extent. A recent study has introduced a layered model to describe mass transport for a FO
112 membrane fouled [33]. This work provides a model to predict the performance of a fouled FO. In
113 turn, their findings have critically mentioned the cake structural parameter (S_c), and the foulant
114 pore hydraulic diameter (D_h) as vital parameters influencing both water and solute transport [33]
115 [34]. However, the profile of CPs in the presence of fouling has not been quantified in these
116 studies. Although an earlier study (with AL-FS configuration) has reported on the couple effects
117 of dilutive ICP and of the fouling occurring on the active layer-facing the feed solution [26], this
118 work did not quantify the effects of CPs and C-ECPs on overall FO performance. In summary,
119 whilst the fouling-enhanced CP and the ICP self-compensation have been noted as having impacts
120 on the overall productivity of the fouled membrane, little attention has been directed towards a
121 quantitative evaluation of the synergistic outcome of CPs and C-ECPs on overall FO performance.

122
123 Given the research gaps indicated above, our current study focused on the following objectives: i)
124 To investigate the CPs and fouling-induced flux decline patterns of the TFC FO membrane under
125 different fluxes for long-term operation of 120h whilst additionally studying the effect of feed
126 compositions (low/high protein fraction present in wastewater) on flux decline and fouling
127 reversibility; and ii) To quantify the effects of CPs (ICP, ECPs), the fouling-induced cake-
128 enhanced CP, and the fouling resistance on the performance of a TFC membrane. Not only does
129 the current study characterize the fouling behaviors by determining morphology, thickness and
130 functional group of fouling layer, but also it covers the calculation of cake structural parameter,
131 and when appropriate the effect of cake-enhanced CP. Finally, correlations between factors i.e.,

132 fouling layer thickness, cake structural parameter, cake-enhanced CP, fouling resistance, and flux
 133 decline rate was explored. This work lays a good foundation for an understanding of transport
 134 phenomena and fouling in the FO process, and also provides a guide for operation of the FO
 135 system.

136 2. Theory

137 A layered model has been developed to describe mass transport for a fouled FO membrane [33].
 138 As foulants are deposited on the active layer surface, this will result in increasing solute
 139 concentration at the interface between the active layer and cake layer; this effect has been labelled
 140 as cake-enhanced CP [29] [32]. Briefly, the effects of ICP, ECPs (concentrative ECP in the feed
 141 side and dilutive ECP in draw side), and cake layer-enhanced CP phenomenon are presented in
 142 Fig. 1.



143
 144 **Figure 1.** Schematic representation of solute concentration profiles across a fouled membrane
 145 under FO mode. ICP, ECPs, and cake-enhanced CP phenomena are accounted for the support

146 layer, boundary layers, and inside the foulant cake layer, respectively. The co-ordinate x is positive
 147 in the direction of positive J_w .

148 **2.1. Internal concentration polarization**

149 Conceptually, the transport of solute in the support layer include two solute fluxes in opposite
 150 directions: a diffusive flux ($D_s dC/dx$) and a convective flux ($J_w C$) [28].

$$151 \quad J_w C + J_s = D_s \frac{dC}{dx} \quad (1)$$

152 where J_s is the reverse draw solute flux. J_w is the water flux, C is solute concentration, and D_s is
 153 an ‘effective’ solute diffusion coefficient for the support layer taking into account the porosity and
 154 tortuosity of the support. As commonly accepted, D_s is calculated from:

$$155 \quad D_s = \frac{D_D \varepsilon}{\tau} \quad (2)$$

156 where D_D is the draw bulk diffusion coefficient, ε is the porosity of the support layer, and τ is the
 157 tortuosity of the support layer.

158 It has been noted that ICP occurring in the support layer decreases water permeation significantly
 159 across the membrane. Under FO mode (AL-FS), dilutive ICP causes a reduction in the osmotic
 160 pressure gradient across the active layer and thus a reduction in the water flux [26]. The boundary
 161 condition in the support layer is defined as Eq. (3).

$$162 \quad \begin{cases} x = 0, C = C_i \\ x = l_{eff}, C = C_{sl} \end{cases} \quad (3)$$

163 where C_i is the solute concentration at the interface between the active layer and support layer, C_{sl}
 164 is the solute concentration on the membrane surface of the support layer, l_{eff} is the effective
 165 thickness of the support layer, which is regarded as the structural parameter of the support layer
 166 (S_s). Hence the mass transfer coefficient in the support layer (k_s) can be calculated as Eq. (4).

$$167 \quad k_s = \frac{D_s}{l_{eff}} = \frac{D_D \varepsilon}{l_{sl} \tau} = \frac{D_D}{S_s} \quad (4)$$

168 From the boundary condition in the support layer in Eq. (3), Eq. (5, 6) are obtained from integrating
 169 Eq. (1). Consequently, the solute concentration at the interface between the support and active
 170 layer (C_i) was defined as below:

$$171 \quad C_i = \exp\left(-\frac{J_w}{k_s}\right)\left(C_{sl} + \frac{J_s}{J_w}\right) - \frac{J_s}{J_w} \quad (5)$$

$$172 \quad C_i = C_{sl} \exp\left(-\frac{J_w}{k_s}\right) - \frac{J_s}{J_w} \left[1 - \exp\left(-\frac{J_w}{k_s}\right)\right] \quad (6)$$

173 where $C_{sl} \exp\left(-\frac{J_w}{k_s}\right)$ is the dilutive concentration of DS due to DS convection and
 174 $\frac{J_s}{J_w} \left[1 - \exp\left(-\frac{J_w}{k_s}\right)\right]$ is the loss of draw solute due to reverse solute diffusion (RSD). Past work has
 175 demonstrated that the ratio of reverse draw solute flux to the water flux (J_s/J_w) can be expressed
 176 as follows [35]:

$$177 \quad \frac{J_s}{J_w} = \frac{B}{AnRT} \quad (7)$$

178 where A is the water permeability coefficient and B is the solute permeation coefficient, n is the
 179 number of dissolved species e.g., 2 for NaCl, R is the ideal gas constant, and T is operating
 180 temperature. This ratio is taken to be a constant coefficient [35].

181 2.2. External concentration polarization

182 The effect of ECP at the membrane surface needs to be considered when developing a model for
 183 the overall performance of the FO membrane [28] [36]. A further decrease in effective osmotic
 184 pressure is attributable to both concentrative feed solution (concentrative ECP) and dilutive draw
 185 solution (dilutive ECP). Using the same differential equation, and applying it to boundary layers
 186 on the draw and feed solutions, two additional equations are obtained.

187 For the draw side, the boundary condition at the interface between the support layer and the bulk
 188 draw solution is expressed below:

$$189 \quad \begin{cases} x = 0, C = C_{D,b} \\ x = \delta_D, C = C_{sl} \end{cases} \quad (8)$$

190 where $C_{D,b}$ is the concentration of solute in the bulk draw solution, δ_D is the thickness of dilutive
 191 ECP in the draw side, and C_{sl} is the solute concentration on the membrane surface of the support

192 layer. As integrated the solute concentration on the membrane surface of the support layer can be
 193 expressed in Eq. (9):

$$194 \quad C_{sl} = C_{D,b} \exp\left(\frac{-J_w}{k_D}\right) - \frac{J_s}{J_w} \left[1 - \exp\left(\frac{-J_w}{k_D}\right)\right] \quad (9)$$

195 where $C_{D,b} \exp\left(-\frac{J_w}{k_D}\right)$ is the dilutive concentration of DS due to DS convection and
 196 $\frac{J_s}{J_w} \left[1 - \exp\left(\frac{-J_w}{k_D}\right)\right]$ is the loss of draw solute due to reverse solute diffusion (RSD), and k_D is the
 197 mass transfer coefficient on the draw side. It is dependent on the hydrodynamic condition on the
 198 membrane surface.

199 Similarly for the feed side, the boundary condition at the interface between the cake layer and the
 200 bulk feed solution is expressed below:

$$201 \quad \begin{cases} x = 0, C = C_{F,b} \\ x = \delta_F, C = C_{F,c} \end{cases} \quad (10)$$

202 where $C_{F,b}$ is the concentration of solute in the bulk feed solution, δ_F is the thickness of the
 203 concentrative ECP on the feed side, $C_{F,c}$ is the solute concentration on the surface of foulant
 204 cake layer. Consequently, the solute concentration on the surface of the foulant cake layer ($C_{F,c}$)
 205 can be expressed as Eq. (11):

$$206 \quad C_{F,c} = C_{F,b} \exp\left(\frac{J_w}{k_F}\right) + \frac{J_s}{J_w} \left[\exp\left(\frac{J_w}{k_F}\right) - 1\right] \quad (11)$$

207 where $C_{F,b} \exp\left(\frac{J_w}{k_F}\right)$ is the concentration of FS due to FS convection and $\frac{J_s}{J_w} \left[\exp\left(\frac{J_w}{k_F}\right) - 1\right]$ is an
 208 accumulation of draw solute due to reverse solute diffusion (RSD), and k_F is the mass transfer
 209 coefficient on the feed side. This depends on the hydrodynamic condition at the membrane surface
 210 on the feed side.

211 The local mass transfer for the hydrodynamic ECP boundary layer of fluids is related to a primary
 212 local Sherwood number. In general terms, prediction on the local mass transfer coefficient on both
 213 the feed side (k_F) and draw side (k_D) can be made from a well-known Graetz equation of
 214 rectangular channel. However, our current channel is relatively short and the membranes surface
 215 has a 'ridge and valley' structure typical of TFC membranes. Both of these affects will make

216 predictions from theory problematic and the Graetz equation is best used to check a directly
 217 calculated value.

218 Combining the equations above, and as reported elsewhere, the water flux in the absence of foulant
 219 can be expressed as [37]:

$$220 \quad J_w = A \left\{ \frac{\pi_{D,b} \exp\left[-J_w \left(\frac{1}{k_D} + \frac{S_s}{D_D}\right)\right] - \pi_{F,b} \exp\left(\frac{J_w}{k_F}\right)}{1 + \frac{B}{J_w} \left\{ \exp\left(\frac{J_w}{k_F}\right) - \exp\left[-J_w \left(\frac{1}{k_D} + \frac{S}{D_D}\right)\right] \right\}} \right\} \quad (12)$$

221 Therefore, we define the external mass transfer coefficients using four actual data sets of fluxes
 222 from baseline experiments with various DS and a foulant free FS. Based on A, B and S determined
 223 before, the methodology of obtaining the best estimates of mass transfer coefficients was to change
 224 the value of k_F and k_D until a sum of square error (SS_{err}) between measured flux and estimated flux
 225 achieve a minimum value. The method has been used previously elsewhere [38]. The outcome is
 226 summarized in Table 1.

227 As using different draw solution concentration (0.75, 1.5 M), it should be noted that change in
 228 external mass transfer coefficients need to be considered. As reported these values could be linked
 229 to the base values, expressed by Eq. 13 [38]:

$$230 \quad \frac{k}{k_o} = \left(\frac{D}{D_o}\right)^{2/3} \quad (13)$$

231

232 **2.3. Cake-enhanced concentration polarization**

233 For a fouled membrane in a FO process it is has been recognized that the build-up of the foulant
 234 cake layer not only causes a hydraulic resistance but there is also hindered diffusion of solutes. As
 235 reported previously [33], the cake layer creates the porous and tortuous structure of foulant and
 236 one can define a cake structural parameter (S_c) consisting of the cake layer thickness (δ_c), tortuosity
 237 (τ_c), and porosity (ε_c), such that $S_c = \delta_c \tau_c / \varepsilon_c$. Boundary conditions were defined as:

$$238 \quad \begin{cases} x = 0, C = C_{F,c} \\ x = S_c, C = C_{al} \end{cases} \quad (14)$$

239 Consequently, the draw solute concentration at the interface between the foulant cake layer and
 240 active layer surface can be expressed by Eq. (15).

$$241 \quad C_{al} = C_{F,c} \exp\left(\frac{J_w S_c}{D_c}\right) + \frac{J_s}{J_w} \left[\exp\left(\frac{J_w S_c}{D_c}\right) - 1 \right] \quad (15)$$

242 where D_c is the diffusion coefficient of solutes in the foulant cake layer. The term D_c/S_c represents
 243 the mass transfer k_c within the foulant cake layer.

244 **2.4. Modeled water flux of a fouled membrane**

245 Based on the solution-diffusion model for the nonporous rejection layer, water flux is defined as
 246 follows:

$$247 \quad J_w = A \Delta\pi_m \quad (16)$$

248 where A is the water permeability coefficient, and $\Delta\pi_m$ is the effective osmotic pressure. However,
 249 the pressure drop induced by the foulant cake layer decreases the effective driving force [33] [39].

250 The pressure drop across the cake layer (ΔP_c) can be expressed as a function of cake structural
 251 parameter (S_c) and pore hydraulic diameter of the cake layer (D_h). Thus,

$$252 \quad \Delta P_c = \frac{32\mu S_c J_w}{D_h^2} \quad (17)$$

253 Therefore, water flux across the active layer should be calculated from:

$$254 \quad J_w = A (\Delta\pi_m - \Delta P_c) = A \left(\Delta\pi_m - \frac{32\mu S_c J_w}{D_h^2} \right) \quad (18)$$

255 where D_h is the hydraulic diameter of the pores in the cake and μ is the viscosity of the solution
 256 within the cake layer.

257 Combining Eq. (6) with Eq. (9), C_i can be expressed as follows:

$$258 \quad C_i = C_{D,b} \exp\left[-J_w \left(\frac{1}{k_D} + \frac{S_s}{D_d}\right)\right] + \frac{J_s}{J_w} \left\{ \exp\left[-J_w \left(\frac{1}{k_D} + \frac{S_s}{D_d}\right)\right] - 1 \right\} \quad (19)$$

259 Combining Eq. (11) with Eq. (15), C_{al} can be expressed as follows:

$$260 \quad C_{al} = C_{F,b} \exp\left[J_w \left(\frac{1}{k_F} + \frac{S_c}{D_c}\right)\right] + \frac{J_s}{J_w} \left\{ \exp\left[J_w \left(\frac{1}{k_F} + \frac{S_c}{D_c}\right)\right] - 1 \right\} \quad (20)$$

261 From Eq. (19) from Eq. (20), the effective concentration difference ($\Delta C_m = C_i - C_{al}$) can be obtained
 262 as follows:

$$\Delta C_m = \frac{C_{D,b} \exp\left[-J_w \left(\frac{1}{k_D} + \frac{S_S}{D_D}\right)\right] - C_{F,b} \exp\left(\frac{J_w}{k_F} + \frac{S_C}{D_C}\right)}{1 + \frac{B}{J_w} \left\{ \exp\left(\frac{J_w}{k_F} + \frac{S_C}{D_C}\right) - \exp\left[-J_w \left(\frac{1}{k_D} + \frac{S_S}{D_D}\right)\right] \right\}} \quad (21)$$

Assuming osmotic pressure is linearly proportion to the salt concentration, then from Eq. (21) and Eq. (18) the modeled water flux of a fouled FO membrane can be calculated using Eq. (22)

$$J_w = A(\Delta\pi_m - \Delta P_c) = A \left\{ \frac{\pi_{D,b} \exp\left[-J_w \left(\frac{1}{k_D} + \frac{S_S}{D_D}\right)\right] - \pi_{F,b} \exp\left(\frac{J_w}{k_F} + \frac{S_C}{D_C}\right)}{1 + \frac{B}{J_w} \left\{ \exp\left(\frac{J_w}{k_F} + \frac{S_C}{D_C}\right) - \exp\left[-J_w \left(\frac{1}{k_D} + \frac{S_S}{D_D}\right)\right] \right\}} - \frac{32\mu S_c J_w}{D_h^2} \right\} \quad (22)$$

3. Materials and methods

3.1. FO membrane and laboratory-scale system

The FO membrane used in the current study, was a polyamide-thin film composite (PA-TFC) cut from a spiral wound FO element (CSM FO-8040, Toray), followed by soaking in deionized (DI) water and then stored at 4 °C. Thickness, morphology, hydrophobicity/hydrophilicity, surface roughness of the pristine membrane was measured. Such intrinsic properties of the TFC membrane i.e., water permeability (A), solute permeability (B), and structural parameter (S) were simultaneously determined using the non-pressurized method with single-stage FO based test. This method has been fully described previously [40]. The specific properties of the membrane are presented in Table S-1.

As illustrated in Fig. S1, the FO cell made of acrylic consisted of two rectangular channels with dimensions of 75 mm (length) × 25 mm (width) × 3 mm (height) and an effective membrane area of 1875 mm². No spacer was used for either the feed or draw channels of the FO cell. An identical flow rate of 300 mL/min (equivalent cross-flow velocity (CFV) of 6.66 cm/s) was maintained in both the feed and draw channels. Operation in counter-current mode was set using the magnetic drive gear pumps. Digital pressure gauges with a precision of 0.005 bar were installed in the DS and FS channels to monitor the maintenance of FO mode. For both the feed solution (FS) and draw solution (DS), the temperature was maintained at 24±1 °C using a water bath. For all experiments, the FO system was operated with an active layer facing the feed solution (AL-FS). An electronic

287 mass balance was used to record the variation in the DS mass and then converted to volume to
288 determine the water flux as Eq. 23.

$$289 \quad J_w = \frac{1}{A_m} \frac{\Delta V}{\Delta t} \quad (23)$$

290 where J_w is the water flux (LMH) at time t , A_m is the effective membrane area (0.001875 m² in the
291 current study), V is the volume of collected permeate (L), and t is the time for collecting the
292 permeate (h).

293 Meanwhile, a conductivity meter was installed into the feed tank to measure the solute mass
294 diffused from draw solution to feed solution. Reverse solute flux (RSF) was calculated as Eq. (24)

$$295 \quad J_s = \frac{C_{F,f}(V_{F,0} - J_w A_m t) - C_{F,0} V_{F,0}}{A_m t} \quad (24)$$

296 where $V_{F,f}$ is the final volume of the feed solution, $V_{F,0}$ is the initial volume of the feed solution,
297 $C_{F,f}$ in the final concentration of the feed solution, and $C_{F,0}$ is the initial concentration of the feed
298 solution.

299 **3.2. Organic foulants**

300 Alginate (ALG), Bovine serum albumin (BSA), Humic acid (HA) were used as adequate models
301 of organic matters to represent for polysaccharides, protein, and humic matter, which are major
302 foulants in wastewater (WW) [41] [42]. They were supplied in powder form (Sigma-Aldrich,
303 USA). The HA stock solution (2 g/L) was prepared by dissolving the powder into 0.01 mol/L
304 NaOH solution (pH > 12) to ensure complete dissolution, followed by filtering through a 1 μm
305 glass fiber membrane to remove residual non-dissolved matter. Meanwhile, the other stock
306 solutions were prepared by dissolving 2 g of the powder in 1 L of DI water. All solutions were
307 stirred over 24 h and then stored at 4°C.

308 **3.3. Fouling experiments**

309 Feed solution for the baseline and fouling tests had a consistent concentration of 10 mM NaCl to
310 avoid changes in ionic strength (IS), causing unintended changes in flux behavior. To simulate
311 real WW the desired foulant concentration was set at 160 mg/L based on findings from a previous

312 study [43]. As denoted in Table 1, fouling was investigated under two scenarios: low protein
313 fraction (ALG: HA: BSA of 2:2:1 wt/wt) and high protein fraction (ALG: HA: BSA of 2:1:2
314 wt/wt). An estimate of a critical flux value was made based upon our previous study [23] and
315 initial fluxes around this value were selected for investigation; values were 24, 30, and 34 LMH
316 (corresponding to DS concentration of 0.75; 1.0; 1.5 M NaCl, respectively). A long-term operation
317 was run over 12 cycles giving 120 h operation in total. To minimize the increase in foulant
318 concentration in FS, after a cycle of 10 hours fresh FS and DS were prepared for the next cycle.
319 For the former fouling scenario, the flux decline patterns were observed under different fluxes. In
320 addition, the effect of feed compositions (low/high protein fraction present in WW) on the fouling
321 reversibility was also evaluated during operation. In detail, for all FO fouling tests a membrane
322 coupon was placed in the test cell and the flow rate of the DS and FS adjusted to achieve the
323 desired CFV value of 6.66 cm/s. Both solutions were prepared with the same volume of 2 L. A
324 stabilization testing for 30 mins was performed using DI water for both the FS and DS sides prior
325 to all experiments. A typical test included the baseline stage followed by the fouling one. A
326 difference between the two stages was indicative of fouling by the foulants in the FS. The flux
327 decline curve in the baseline test was used to normalize the flux decline curves in the fouling test,
328 thereby obtaining a measure of the flux decline due to fouling alone. This normalization has been
329 described in detail elsewhere [17].

330 **Table 1.** Feed solution composition

Feed solution	Foulant concentration (mg/L)	Ionic strength (mM)
Baseline test	-	10
Fouling test with Simulated wastewater		
ALG: HA: BSA (2:2:1 wt/wt) (low protein fraction)	160	10
ALG: HA: BSA (2:1:2 wt/wt) (high protein fraction)	160	

331
332 Examination of fouling reversibility and membrane autopsies were performed. For calculating
333 fouling reversibility (FR), physical flushing on the feed side using DI water was immediately

334 performed for 15 mins with an increased CFV value of 13.32 cm/s, followed by a measurement of
 335 flux at the initial CFV. In addition, when flux declines by over 15%, chemical cleaning was
 336 adopted to recover flux. The conditions for chemical cleaning were 1% NaOH, at a feed flow rate
 337 of 600 mL/min (CFV of 13.32 cm/s), cleaning time for 15 mins, followed by DI flushing for 15
 338 mins to remove residual NaOH. Flux recovery or fouling reversibility (FR) in FO was calculated
 339 as follows [21].

$$340 \quad FR (\%) = \frac{J_{xr} - J_{xf}}{J_{xi} - J_{xf}} \times 100 \quad (25)$$

341 where J_{xi} is the water flux of a virgin membrane, J_{xf} is water flux after fouling (LMH), and J_{xr} is
 342 the water flux after physical or chemical cleaning (LMH).

343 **3.4. Quantification of CPs and the cake foulant-enhanced CP**

344 Based on theory, the loss of driving force is attributed to the effects of ICP in the support layer,
 345 ECPs (concentrative ECP in the feed side, and dilutive ECP on the draw side) and additional cake
 346 layer-ECP phenomena. These effects can be quantified by Eq. (26), Eq. (27), Eq. (28), and Eq.
 347 (29) respectively:

$$348 \quad \text{Dilutive ICP} = \frac{C_{sl} - C_i}{C_{D,b} - C_{F,b}} \times 100 \quad (26)$$

$$349 \quad \text{Dilutive ECP} = \frac{C_{D,b} - C_{sl}}{C_{D,b} - C_{F,b}} \times 100 \quad (27)$$

$$350 \quad \text{Concentrative ECP} = \frac{C_{F,c} - C_{F,b}}{C_{D,b} - C_{F,b}} \times 100 \quad (28)$$

$$351 \quad \text{Cake 'foulant enhanced' CP} = \frac{C_{al} - C_{F,c}}{C_{D,b} - C_{F,b}} \times 100 \quad (29)$$

352 **3.5. Cake structural parameter and fouling resistance determination**

353 Cake structural parameter (S_c) of the foulant layer obtained from the conditions of different fluxes,
 354 various feed compositions were calculated for comparison. In detail, the cake structural parameter
 355 is predicted in an iterative process by minimizing the error between calculated flux ($J_{w, cal}$) and
 356 experimental flux ($J_{w, measured}$). Operating conditions (Table 2) and model parameters (Table 3)

357 were used to calculate the cake structural parameter. A detailed calculation procedure is shown in
 358 Fig. 2.

359

360 **Table 2.** Operating conditions of membrane channel

Condition	Description	Unit	Value
L	Length of membrane channel	m	0.075
W	Width of membrane channel	m	0.025
d	Height of membrane channel	m	0.003
T	Temperature	K	293
CFV	Cross-flow velocity (Both feed and draw sides)	m/s	0.067

361

362 **Table 3.** Model parameters used for calculating cake structural parameter

Parameter	Description	Unit	Value
A	Water permeability coefficient	$\text{m}^3/\text{m}^2 \cdot \text{s} \cdot \text{Pa}$	9.14×10^{-12} (This study)
		LMH Bar ⁻¹	3.29 (This study)
B	Solute permeability coefficient	$\text{m}^3/\text{m}^2 \cdot \text{s}$	4.75×10^{-7} (This study)
		LMH	1.71 (This study)
S_s	Structural parameter of the support layer	m	2.10×10^{-4} (This study)
D_D	Bulk diffusion coefficient of draw solution	m^2/s	1.38×10^{-9} (1.5 M NaCl) [44]
			1.41×10^{-9} (1.0 M NaCl) [44]
			1.43×10^{-9} (0.75 M NaCl) [44]
D_F	Bulk diffusion coefficient of feed solution	m^2/s	1.62×10^{-9} [44]
D_C	Diffusion coefficient in foulant cake layer	m^2/s	1.29×10^{-9} [33]
μ	Dynamic viscosity of water	kg/m.s	8.9×10^{-4}
ρ	Density of water	kg/m ³	998
D_h	The hydraulic pore diameter of foulants	m	2.58×10^{-7} (This study)

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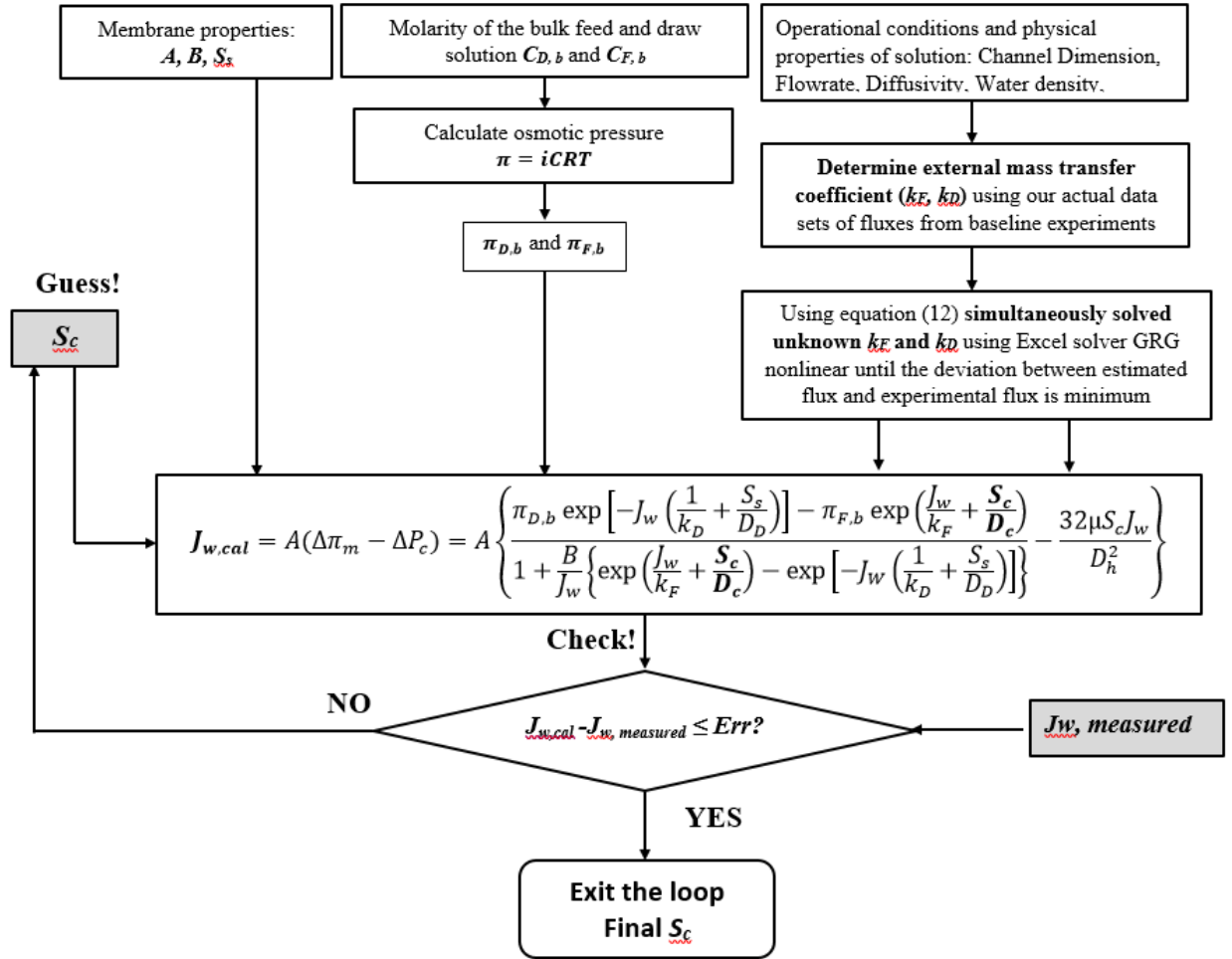
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373 **Figure 2.** Flow chart of the calculation procedure of cake structural parameter (S_c)

374 Membrane resistance (R_m) and resistance due purely to concentration polarization (R_{cp}) were
 375 determined using Eq. (33), which was derived based on the osmotic-resistance for osmotically
 376 driven membrane process [45]:

$$377 \quad J_w = \frac{\text{Driving force}}{\mu(R_m + R_{cp})} \quad (33)$$

378 When fouling occurs, total resistance (R_t) will impact water flux and is calculated in Eq. (34). This
 379 resistance includes membrane resistance (R_m), fouling resistance (R_f), and resistance due purely to
 380 concentration polarization (R_{cp})

$$381 \quad R_t = \frac{\text{Driving force}}{\mu J_{w,f}} \quad (34)$$

382 And the foulant resistance R_f is then calculated by subtracting the total resistance R_t with the
 383 pristine membrane resistance R_m and resistance of concentration polarization R_{cp} as follow:

$$384 \quad R_f = R_t - R_m - R_{cp} \quad (35)$$

385 **3.6. Foulant characterization**

386 The morphology (surface and thickness) and composition (e.g., functional groups) of the fouling
387 layer were characterized by Field Emission Scanning Electron Microscopes (FE-SEM) and the
388 Attenuated Total Reflectance-Fourier Transform Infrared Spectroscopy (ATR-FTIR),
389 respectively. In detail, surface and cross-sectional images of morphological features (virgin,
390 fouled, and cleaned FO membranes) were observed using a Field Emission Scanning Electron
391 Microscopes (FE-SEM) (S-4700, Hitachi, Japan) to examine the presence and the morphology of
392 foulant on the membrane surface. Prior to SEM analysis, all samples were air-dried in a desiccator
393 overnight and subsequently coated with platinum for 3 min. Spectra of samples were obtained
394 using Attenuated Total Reflectance-Fourier Transform Infrared (ATR-FTIR) Spectroscopy (Perkin-
395 Elmer IR 2000 Series, Billerica, MA, USA) with a resolution of 4 cm^{-1} in the range of 4000 cm^{-1}
396 to 700 cm^{-1} . These measurements were used to indicate the organic functional groups of the
397 pristine, fouled, and cleaned membrane. All samples were air-dried in a desiccator overnight
398 before analysis. The aggregation size of foulants was measured by dynamic light scattering.

399

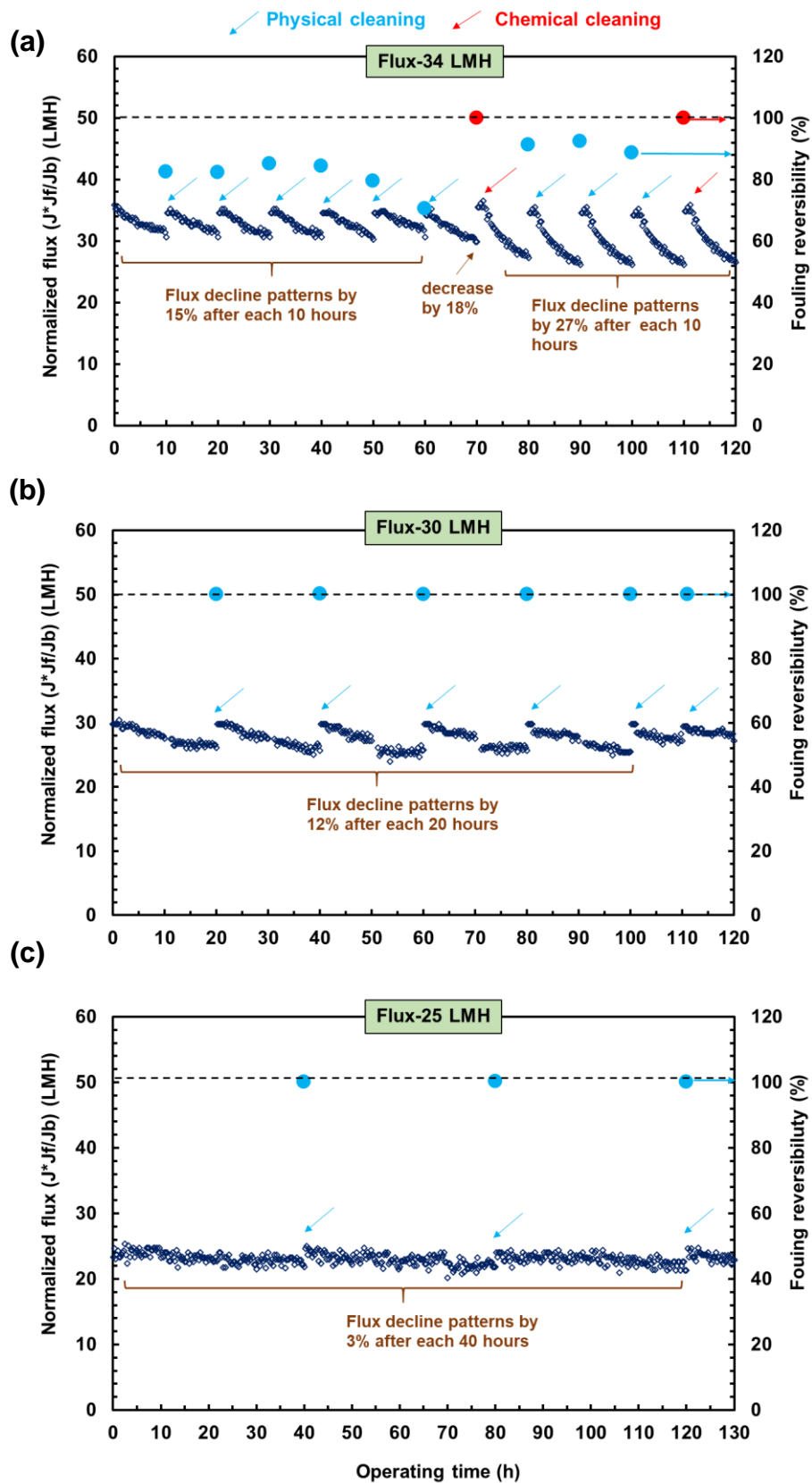
400 **4. Results and discussion**

401 **4.1. Flux decline patterns and the fouling reversibility under operating conditions**

402 Based on a critical value for complex fouling (30 LMH) obtained from our past work [23], a long-
403 term 120 h investigation into different fluxes, at and either side of 30 LMH, was carried out. The
404 water flux was normalized to indicate the extent of fouling. As denoted in Fig. 3a, for the condition
405 of 34 LMH a flux declined of 15% was observed in a 10 h operation after which there was a
406 physical clean. This trend was maintained for the first six cycles, i.e. 60 h operation. As observed
407 in Fig. S3-a1, b1, a cake layer with a thickness of $4.78 \pm 0.32\ \mu\text{m}$ was formed on the membrane
408 surface. This layer made up an increased portion of total resistance and caused cake-ECP; thus this
409 was the cause of the significant flux decline. As fouling became severe, physical cleaning was
410 conducted to recovery the initial flux but it could only attain low fouling reversibility; FR was 82-

411 85% during 60 h operation. This can be attributed to the fact that operation was above the critical
412 value previously identified [16] [23]. During the first 50 h, the FR decreased from 85% to 79%.
413 Due to foulant accumulation on the membrane surface higher decreases in water flux were
414 observed in subsequent cycles. Direct observations indicated that foulants still remained on the
415 membrane surface after imposing physical cleaning (Fig S3-c1, d1). Under this relatively high flux
416 (34 LMH) the fouling layer becomes denser and thicker with time (Fig S3-a1, b1, and a2, b2)
417 which renders the fouling layer less sensitive to changes in hydrodynamic conditions i.e.,
418 increasing CFV. The result is irreversible fouling. To recover the membrane a chemical cleaning
419 was conducted after 70 h. As expected the water flux was fully restored (100% FR). However, this
420 action not only increased the initial flux to 37 LMH but also altered the flux decline pattern which
421 became more severe; there was a decrease of 27% after each subsequent 10 h operation. This
422 implies that the PA-TFC membrane surface could be modified by chemical agents, which resulted
423 in a change in water flux behavior. As reported, chemical cleaning agents caused changes in
424 membrane properties [46]. The pore size of TFC could be enlarged after alkaline exposure and this
425 is attributed to the increased electrostatic interactions among the deprotonated carboxylic
426 functional groups of the polyamide active skin layer [47]. Since the FR continued to reduce from
427 93% to 88% after 100 h of operation, a second chemical cleaning was made. At 27% the
428 subsequent flux decline remained the same.

429 To minimize the degree of fouling the study turned to lower fluxes and the frequency of physical
430 cleanings was reduced as shown in Fig. 3. For 30 LMH a flux decline of 12% was observed over
431 20 h operation. For 25 LMH the decline was 3% over 40 h operation, indicating very minor fouling.
432 This condition appeared below critical values, which facilitated a sustainable operation and
433 cleaning minimization [48]. It is important to note that for both 30 and 25 LMH water flux was
434 restored fully by a physical cleaning. To further elucidate the different flux decline patterns, the
435 influence of CPs (ECPs, ICP, and Cake-ECP) and the fouling resistance were determined. The
436 results for the different fluxes are presented later in section 4.3.



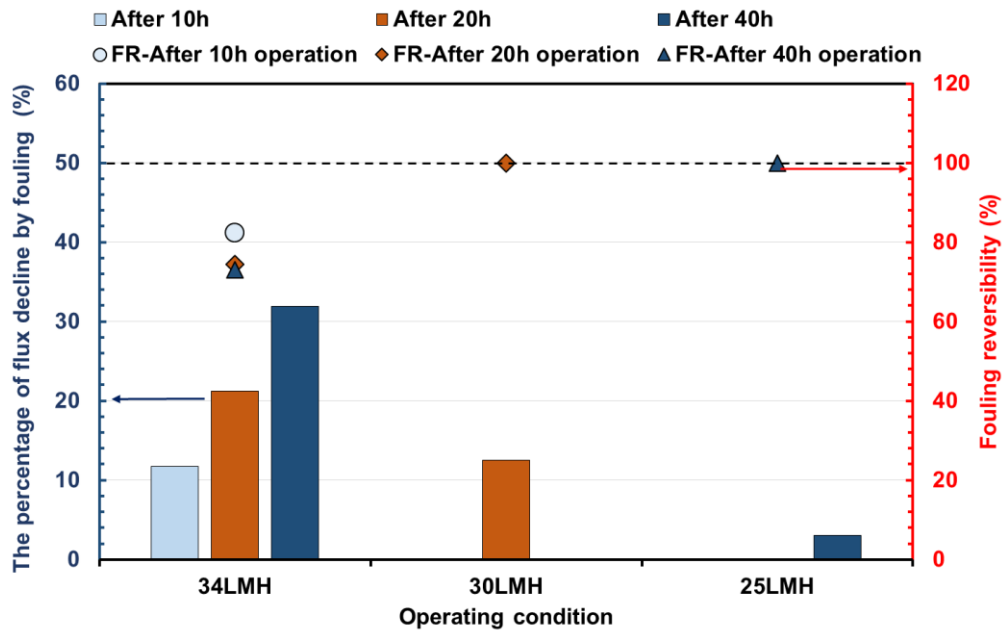
437
 438 **Figure 3.** Flux decline patterns and the fouling reversibility under different fluxes: (a) 34 LMH;
 439 (b) 30 LMH and (c) 25 LMH. All experiments were conducted at the long-term operation of 120-
 440 130 h. $J_o \times J_f/J_b$ is a normalized representation of the extent of membrane fouling. J_o represents the

441 initial flux, J_f is the flux in the fouling test, J_b is the flux in the baseline flux. Simulated wastewater
442 consists of complex foulants of ALG, HA, and BSA (ALG: HA: BSA= 2:2:1 wt/wt e.g., low
443 protein fraction) and the foulant concentration of 160 mg/L.

444 An analysis of the influence of flux upon FR and flux decline is shown in Fig. 4. As severe fouling
445 occurred at a flux of 34 LMH, the flux declines of 13%; 25% and 32% were observed in 10 h, 20
446 h and 40 h operation. The corresponding fouling layer thicknesses were $4.78 \pm 0.32 \mu\text{m}$, $10.96 \pm$
447 $0.82 \mu\text{m}$, $20.05 \pm 0.52 \mu\text{m}$ respectively (Fig S3- b1, b2, b3). It is generally accepted that a lower
448 % flux decline facilitated higher FR. Although the same flux decline was roughly 12% after 10h
449 and 20 h operation observed for 30 and 34 LMH fluxes, respectively, the former (FR 100%)
450 exhibited a higher FR than the latter (FR 82%). Furthermore full restoration was achieved at 30
451 LMH using simple physical cleaning. This value corresponds with the critical value determined in
452 our past works [16] [23].

453 For the case of 34 LMH, as the extent of FR via physical cleaning decreased from 82% to 73% the
454 degree of flux decline increased from 13 to 21%. Interestingly, there was no significant difference
455 in FRs once physical cleaning was implemented after 20 h (FR 74%) and 40 h (FR 73%), as shown
456 in Figure 4. After cleaning visualization of the outcomes was made by SEM images indicated that
457 a comparable thickness of cake layer i.e., $0.82 \pm 0.07 \mu\text{m}$ and $0.91 \pm 0.15 \mu\text{m}$ were for the cases
458 of 20 h and 40 h operation, respectively (Fig S3-d2, d3). This suggests that for fouling by WW
459 possessing a low protein fraction, the upper cake layer could be readily removed using simple
460 physical cleaning and that the core underlying layer changes slowly.

461 In the literature, the fraction of polysaccharide, protein, and humic acid in WW depends on the
462 treatment stage and technology [41]. Therefore it was of interest to investigate the effect of feed
463 compositions (low protein and high protein fractions) on the FO performance see next section.



464
 465 **Figure 4.** Comparison of the extent of flux decline and fouling reversibility (FR) under different
 466 fluxes (34; 30; 25 LMH). A physical cleaning was conducted for scenarios: after 10 h, 20 h, or 40
 467 h depending on the initial flux. Condition of physical cleaning: DI water at CFV of 13.2 cm/s for
 468 15 mins.

469
 470 **4.2. Effect of feed compositions on % flux decline and the fouling reversibility**
 471 To investigate the influence of the level of the protein fractions present in WW, a flux of 30 LMH
 472 was chosen and the % flux decline and FR were assessed over 100 h operation (Fig. 5-a). For both
 473 levels of protein, the flux decline gradually increased, possibly towards a plateau. This is attributed
 474 to a build-up of irreversible fouling. For just the initial stage of 20 h, the extent of flux decline was
 475 comparable for the two levels of protein, with an average decline of around 12%. This was
 476 confirmed by surface and cross-sectional analyses of the fouled membrane (Fig. 5-b, c, d); after
 477 operated in 20 h the cake layer thicknesses were $2.51 \pm 0.25 \mu\text{m}$ and $2.90 \pm 0.20 \mu\text{m}$ for the high
 478 and low protein cases, respectively. Thereafter, the fouling extent became more serious for the
 479 high level of protein, with flux decline reaching 18% after 100 h operation, compared with 14%
 480 for the lower level of protein. These results can probably be attributed to different cleaning-
 481 efficiency from two scenarios. Whilst the water flux was fully restored (FR 100%) for WW

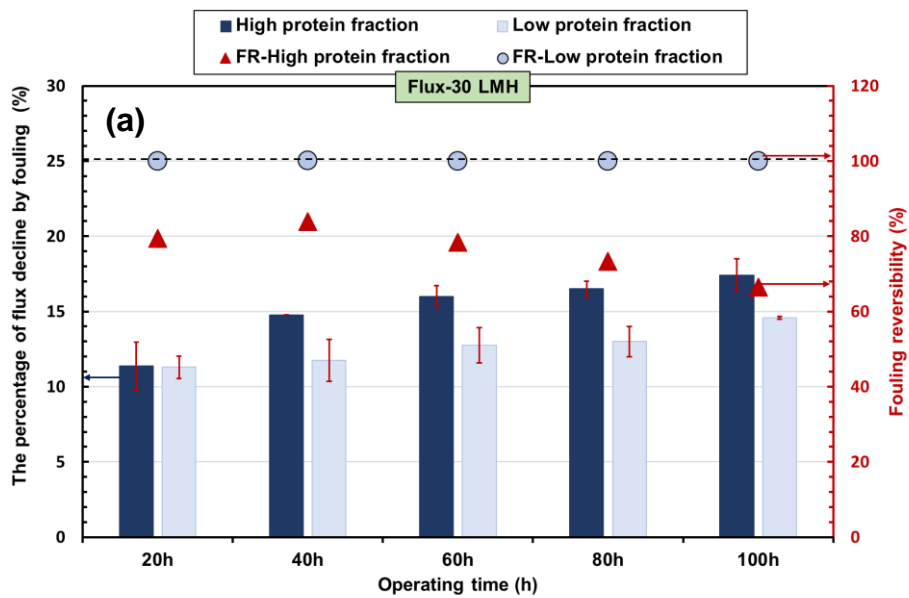
482 possessing low protein fouling, this was not expected for high protein fouling in which the FR was
483 low (67%) after 100 h operation. This fact was confirmed through measuring the cake layer
484 thicknesses after cleaning. As shown in Fig. 5-d, g, values of $0.46 \pm 0.05 \mu\text{m}$ were found for low
485 protein fraction compared with $1.39 \pm 0.20 \mu\text{m}$ for high protein fraction. It is noted that the reverse
486 solute flux of the fouled membrane i.e., 7.4-9.5 $\text{g}/\text{m}^2\cdot\text{h}$ was lower than that of the pristine
487 membrane i.e., 11.9 $\text{g}/\text{m}^2\cdot\text{h}$ (Fig. 6-a). This indicates that the reverse solute flux decreased with the
488 development of the cake layer.

489 As previously reported the formation of complex organic fouling layer rendered the membrane
490 more negatively charged [5] [23]. This increase could reduce the transport of draw solution ions
491 in the reverse direction. Likely, the reverse solute flux of Cl^- ion was hampered by an enhanced
492 electrostatic interaction with the more negatively organic fouling layer. As a final point on the
493 RSF, it is noted that membrane fouled by low protein fraction yielded a higher RSF than that by
494 high protein fraction and this can be probably attributed to different foulant structures e.g., cake
495 structural parameter. Detailed calculation on this parameter is covered in section 4.3.2.

496 The FTIR analyses of the pristine membrane, the fouled membrane, and a cleaned membrane are
497 presented in Fig. 6-b, c. These results indicated that the PA-TFC membrane comprised of a
498 polysulfone layer and polyamide layer. Operation at either low or high protein conditions led to
499 the formation of a cake layer on the membrane surface. Consequently, the peaks of the active layer
500 and support layer of the pristine membrane were attenuated. The results for the fouled membranes,
501 show the emergence of peaks. For the case of low protein fraction, peaks associated with quinone
502 and ketone C-O bond, or aromatic rings (C=C) were found indicating the presence of HA in the
503 top layer of the cake layer. For the high protein fraction the existence of protein -like substance
504 was found. The results showed peaks such as amide A (N-H stretching), amide I (C=O stretching),
505 amide II (N-H in-plane), amide III (C-N stretching), and CH_2 bending at 3280, 1643, 1532, and
506 1450, 1392 cm^{-1} , respectively. Since membrane fouled by low protein fraction could be restored
507 readily, the FTIR spectra obtained from the cleaned membrane are likely to be similar to the

508 pristine membrane. Although a physical cleaning was implemented for membrane fouled by high
509 protein fraction, the results of FTIR showed the presence of a protein-like substance on the cleaned
510 membrane. Overall the findings indicate that protein is a persistent foulant and less sensitive to
511 physical cleaning.

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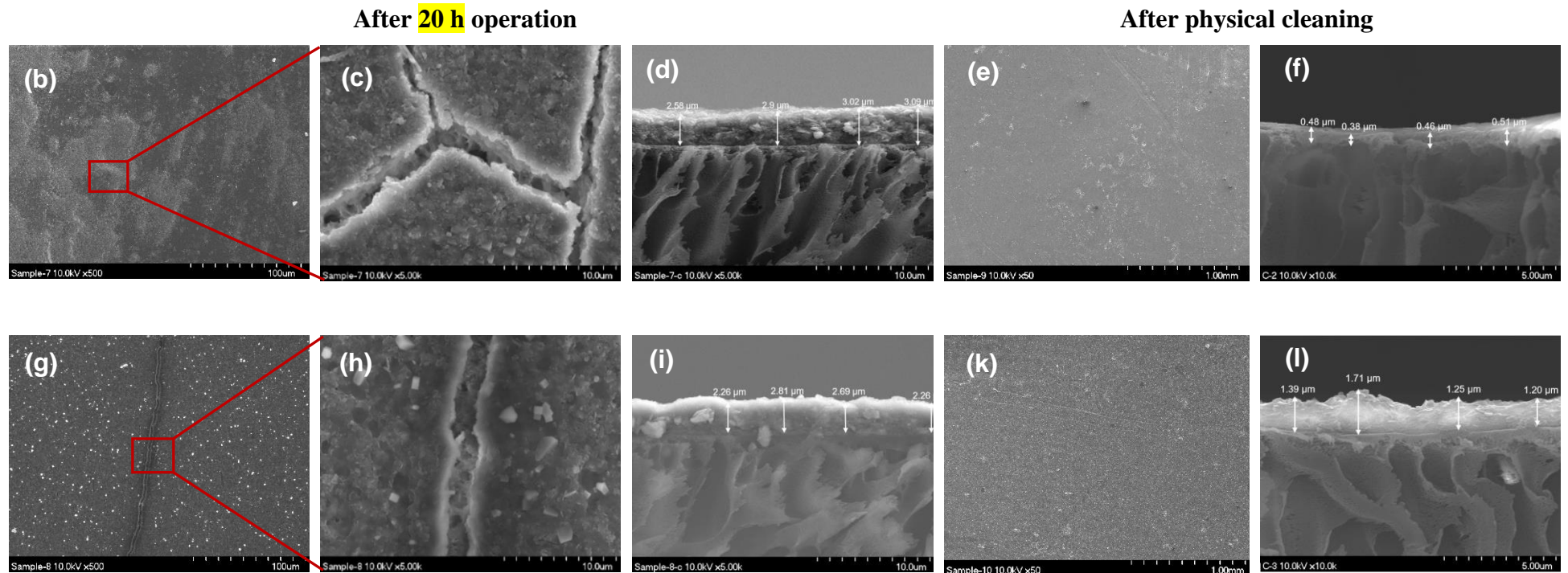
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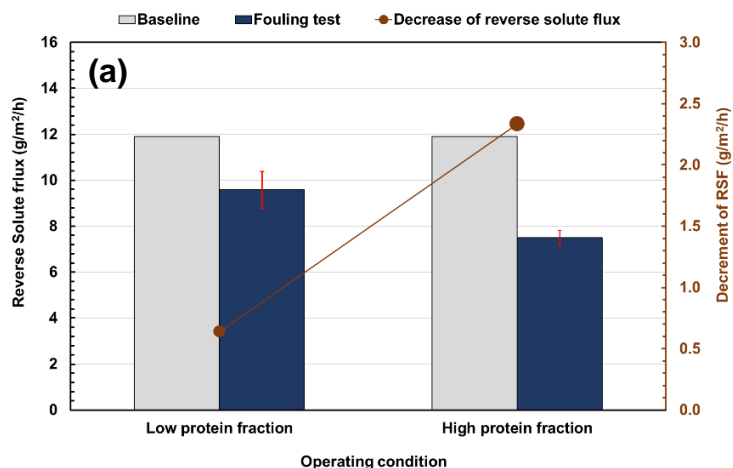
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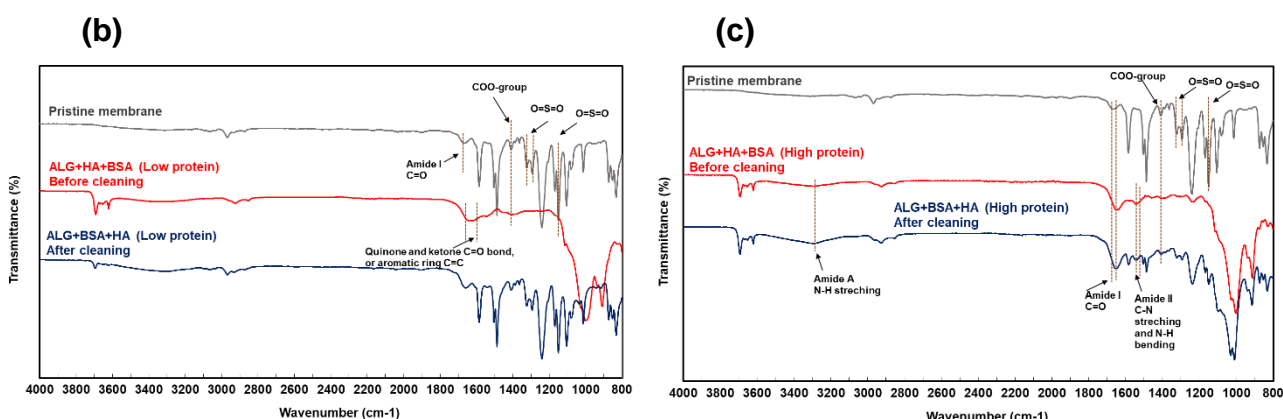


526 **Figure 5.** Comparison of % flux decline and the fouling reversibility under various feed compositions: low protein fraction (ALG: HA: BSA=2:2:1
 527 wt/wt) and high protein fraction (ALG: HA: BSA=2:1:2 wt/wt) (a). All experiments had same flux of 30 LMH. A physical cleaning was conducted after
 528 each 20 h operation to define the fouling reversibility. Typical condition of physical cleaning: DI water with CFV of 13.2 cm/s for 15 minutes. SEM
 529 observation was conducted after 20 h fouling operation and after physical cleaning: low protein fraction (b, c, d, e, f) and high protein fraction (g, h, i, k,
 530 **D**)

531



532



533

534 **Figure 6.** Comparison of reverse solute flux and FTIR measurements as operated at various feed
535 compositions: (low protein of ALG: HA: BSA=2:2:1 wt/wt and high protein fraction of ALG: HA:
536 BSA=2:1:2 wt/wt). All experiments were operated under the same flux of 30 LMH and foulant
537 concentration of 160 mg/L. A reverse solute flux of pristine membrane (baseline) and fouled
538 membrane (fouling test) after 100 h operation (a). FTIR measurements of the pristine membrane,
539 fouled membrane, and cleaned membrane for the cases: low protein fraction (b) and high protein
540 fraction (c).

541

542 4.3. Quantitative analysis of concentration polarization and fouling resistance

543 4.3.1. Different initial fluxes

544 In the FO membrane process, changes in water flux are the result not only of fouling but are also
545 the result of consequential changes in CPs effects. Therefore, to obtain an insightful understanding

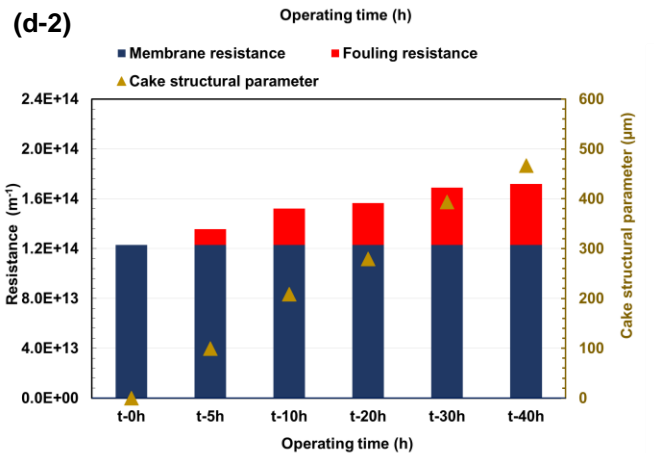
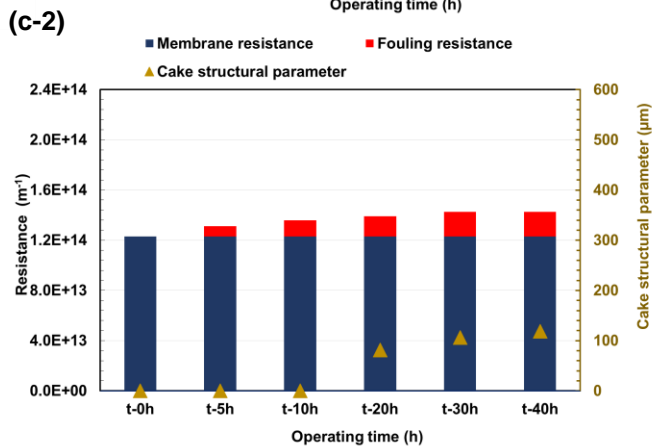
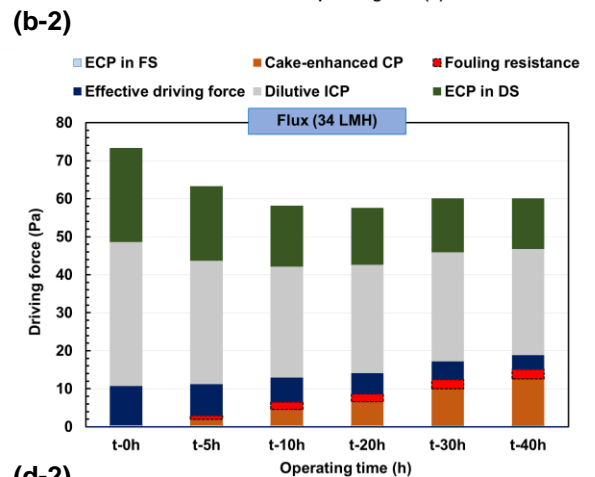
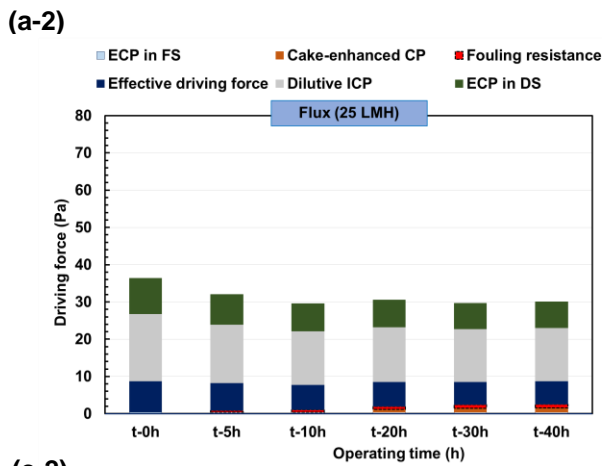
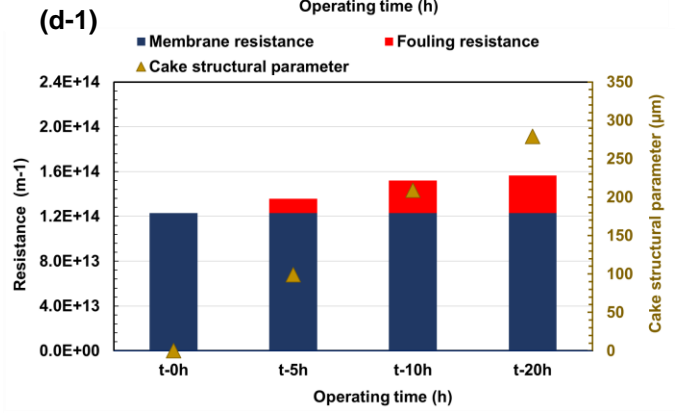
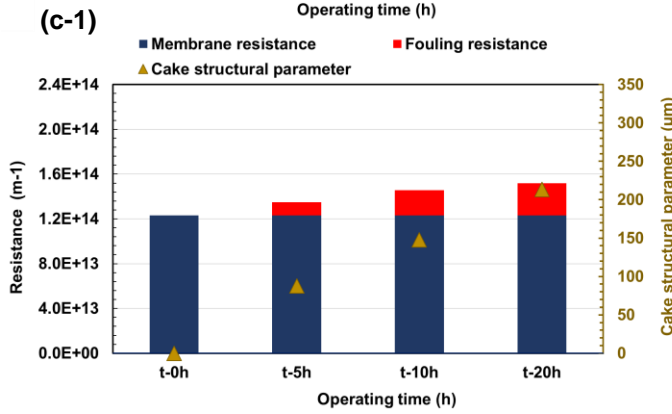
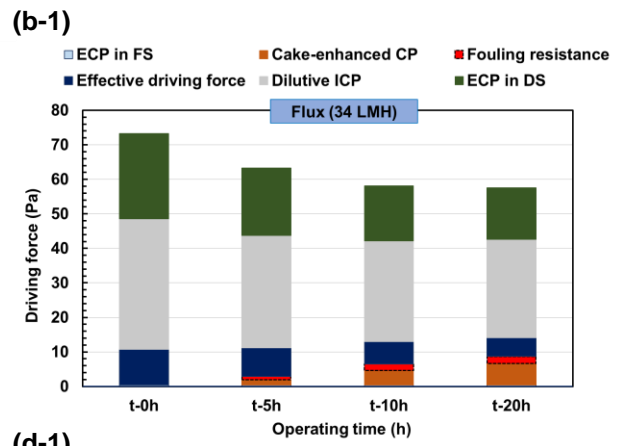
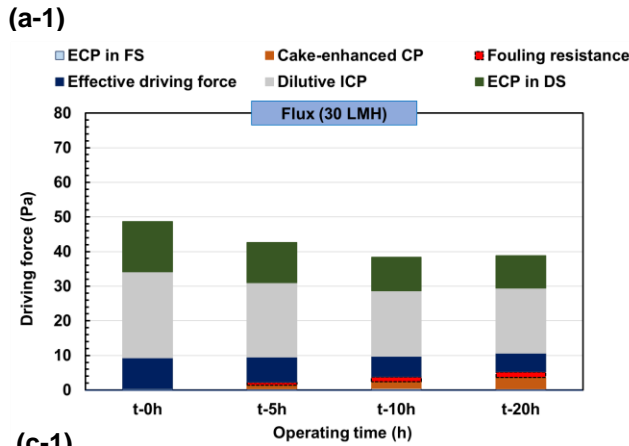
546 of the CPs and fouling effects, we quantified the extent of CPs in each of the transport layers, the
547 fouling resistance of the foulant layer, for operation under different fluxes (Fig. 7).
548 It is generally accepted that higher overall driving force is attained at higher DS concentration and
549 thus exhibits higher initial flux. Such initial effective driving forces were 8.24, 8.95 and 10.31 bar
550 for operational fluxes of 25, 30, and 34 LMH, respectively (Fig. 7-a1, a2, b1 and b2). Whilst the
551 effect of concentrative ECP in FS was very minor due to the low salt concentration (10mM NaCl),
552 the dilutive ECP in DS and ICP had profound effects on the FO performance. These CP effects
553 contributed 76.0%, 80.8%, and 85.3% on overall driving force under fluxes of 25, 30, and 34 LMH
554 respectively (Fig S-5, 6, 7). After 20h operation the water flux declined by fouling was 12% and
555 25% for the fluxes of 30 and 34 LMH respectively and this is attributed to both decrease of the
556 effective driving force (Fig. 7-a1, b-1) and increase in the fouling resistance (Fig. 7-c1, d1). As
557 denoted in Fig. 7a-1, b-1, the cake layer-enhanced CP gradually increased in 20h operation and
558 thus partially contributed to a decrease of the effective driving force. Operation at 30 LMH caused
559 a relatively moderate decrease of the effective driving force from 8.95 to 5.45 bar whereas at a
560 higher flux (34 LMH) it more that halved. As the cake layer evolved and flux reduced, the ICP
561 effect diminished during the 20 h fouling test for both fluxes, and this is phenomenon has been
562 labelled ICP-compensation [49]. As indicated by the results from SEM observation (Fig S4-b1,
563 d1), after 20 h operation, the foulant layer thicknesses were $10.96 \pm 0.82 \mu\text{m}$ and 2.90 ± 0.20 for
564 34 and 30 LMH respectively. Different thicknesses resulted in different fouling resistances and
565 cake structural parameters. The corresponding values for 34 LMH were $3.37 \times 10^{13} \text{ m}^{-1}$ and 279
566 μm , which are higher than those for 30 LMH which were $2.88 \times 10^{13} \text{ m}^{-1}$ and 213 μm). Overall the
567 results indicated that the flux chosen plays a pivotal role in fouling control.

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575 **Figure 7.** Change of CPs profile, cake structural parameter and fouling resistance with time under
 576 different fluxes: 30 LMH (a-1, c-1) vs 34 LMH (b-1, d-1) for a 20 h fouling test conducted without

577 physical cleaning; 25 LMH (a-2, c-2) vs 34 LMH (b-2, d-2) for a 40 h fouling test conducted
578 without physical cleaning.

579 As discussed in section 4.1, negligible fouling i.e., 3% flux decline in 40 h operation, was noticed
580 for operation at 25 LMH, which is below the critical value. For 25 LMH the fouling resistance
581 increased only marginally in 40 h operation (Fig. 7-c2), indicated that the impact of the cake layer
582 on permeate is insignificant and this is due to a thin fouling layer ($1.08 \pm 0.23 \mu\text{m}$) formed on the
583 membrane surface. After 40 h operation, a minor influence of cake layer-enhanced CP and fouling
584 resistance, respectively being 4.1% and 3.4% of the overall driving force (Fig S6-d), was observed.
585 The cake had a small structural parameter ($119 \mu\text{m}$). By way of contrast, a negative impact was
586 apparent at a high flux of 34 LMH where the flux decline led to significant decreases in the
587 effective driving force. Compared to 25 LMH, a thicker cake layer of $20.05 \pm 0.52 \mu\text{m}$ (Fig S4-
588 b2) formed on the membrane surface resulting in a 2.5 times increase in the fouling resistance
589 ($4.88 \times 10^{13} \text{ m}^{-1}$) and a four-fold increase in the cake structural parameter ($466 \mu\text{m}$). The cake layer
590 not only posed an additional resistance to water transport but also enhanced solute concentration
591 near the membrane surface; thereby reducing further the water flux. Although the ICP was offset
592 by the cake layer-induced flux decline, the cake layer-enhanced CP was noticeable with it
593 consuming 20.4% of the total driving force after 40 h operation (Fig S-6c). These findings quantify
594 the significance of the cake layer-enhanced CP under high flux operation.

595

596 **4.3.2 Different feed compositions**

597 It was found that the feed composition possessing high protein fraction triggered a higher flux
598 decline rate and lower FR compared to low protein fraction, as shown in Fig. 5-a. To elucidate the
599 results given in section 4.2, the CPs effect and the fouling resistance from these fouling cases were
600 quantified. Generally, the extent of ICP, ECPs in a membrane fouled by low protein fraction was
601 comparable with that by high protein fraction and this is attributed to operation at the same flux
602 (30 LMH) and similar flux decline, particularly in the initial stage. For the initial 10 h operation

603 the flux decline rate was similar for all cases and this fact is confirmed by calculation on the fouling
604 resistance and the cake structural parameter. As presented in Fig. 8, the corresponding values for
605 fouling resistance and cake structural parameters being $2.28 \times 10^{13} \text{ m}^{-1}$ and $148 \mu\text{m}$, and 2.24×10^{13}
606 m^{-1} and $150 \mu\text{m}$ for low and high protein fraction respectively. Clearly there is no significant
607 difference. However, it is noted that after 100 h operation the fouling resistance ($3.37 \times 10^{13} \text{ m}^{-1}$)
608 and cake structural parameter ($308 \mu\text{m}$) of membrane fouled by high protein fraction were higher
609 than those of low protein fraction i.e., $3.04 \times 10^{13} \text{ m}^{-1}$ and $265 \mu\text{m}$. This leads to a greater extent of
610 C-ECP for the former condition i.e., 4.96 bar, compared to the latter condition i.e., 4.30 bar. Such
611 results confirmed that a higher fouling rate occurred in the conditions of high protein fraction.
612 Generally, the cake structural parameter and the fouling resistance increased gradually for both
613 conditions. This fact implies that physical cleaning after each 20 h operation did not prevent an
614 accumulation of foulant on the membrane surface. The resultant cake layers promoted the effect
615 of C-ECP which was found to account for 10.9-13% of the total driving force.
616 As indicated by the result of cross-sectional analyses (Fig. 5c, d) the fouling layer of the low
617 protein fraction was slightly thicker than that of the high protein fraction. However, the latter
618 generates the higher fouling resistance and cake structural parameter. Such results imply that the
619 high protein fraction present in WW could lead to alterations in the foulant layer structure where
620 protein might strongly interact with alginate and humic acid in a complex fouling solution [23].
621 Overall these findings suggest that the fouling resistance and the cake structural parameter (rather
622 than thickness itself) are the pivotal factors in an evaluation of the extent of fouling.

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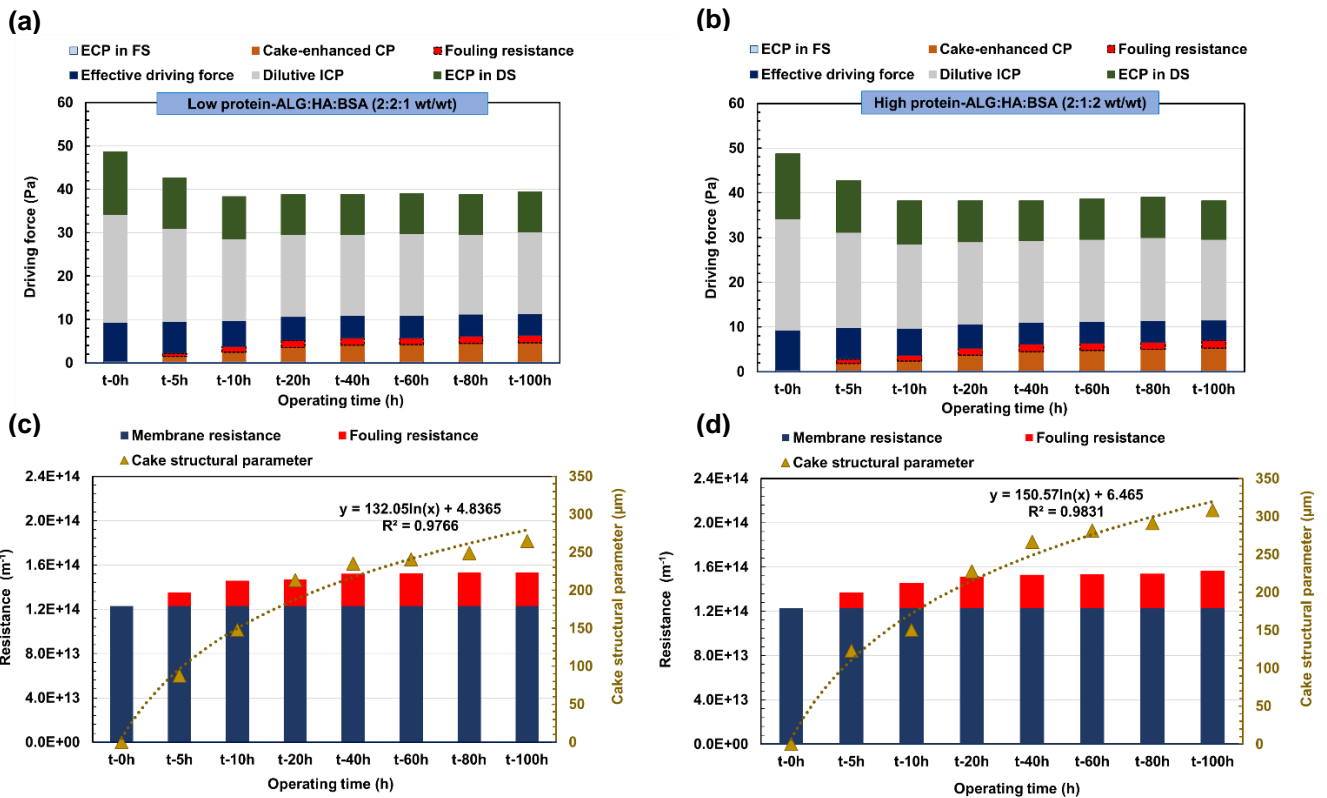
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631 **Figure 8.** Change of CPs profile, cake structural parameter, and fouling resistance with different
 632 feed compositions: Low protein (a, c), high protein (b, d). A long-term fouling test was conducted
 633 for 100 h. Physical cleaning was implemented after each 20 h operation. All experiments were
 634 operated under the same flux of 30 LMH and the foulant concentration of 160 mg/L.

635

636 4.4. Correlation between factors: Cake-enhanced concentration polarization, cake structural 637 parameter, fouling layer thickness, fouling resistance and % flux decline by fouling

638 This section provides an exploration of the correlation coefficients between factors i.e., cake-
 639 enhanced CP, fouling layer thickness, cake structural parameter, and fouling resistance and the
 640 flux decline rate. Table S2 presents such values of these factors under different fluxes (low,
 641 moderate, and high level), fouling time, and various foulant types (single and complex fouling).
 642 Such relationships were evaluated using Pearson's correlation coefficient (Table 2). As expected
 643 the flux decline rate showed a good correlation with the following factors: fouling layer thickness

644 (R= 0.897), fouling resistance (R= 0.908), cake structural parameter (R= 0.925), and cake-
645 enhanced CP (R=0.904).

646 The cake structural parameter showed proficient correlations with cake-enhanced CP (R= 0.986),
647 the fouling resistance (R=0.873). However, this was not attained between cake-enhanced CP and
648 fouling layer thickness as their correlation coefficient was modest (0.794). These results can be
649 probably attributed to different cake structures formed by either single or complex fouling. As
650 previously reported, the tortuosity of an organic foulant cake layer was probably close to 1 [33].
651 If the same flux of 34 LMH was implemented (Table S-2), different cake structural parameters
652 obtained from fouling scenarios could be due to different thickness and porosity of cake layers.
653 Such findings suggested that the feed composition could induce a change in a cake layer-enhanced
654 CP, fouling layer thickness, and thus pose different fouling behaviors.

655 As flux decline by fouling has been partially ascribed to cake layer resistance and the layer
656 thickness, an examination to define the quantitative regression between the fouling layer thickness
657 and the flux decline rate was made. As shown in Fig. 9, a single linear regression and logarithmic
658 model were used for comparison. The latter model showed a stronger correlation and this was
659 observed for the simulated WW with complex foulant ($R^2= 0.9956$). Notably, correlated values
660 for simulated WW with single foulant deviated from the logarithmic regression. These findings
661 revealed that for the feeds used the flux decline rate and fouling layer thickness can be linked
662 through logarithmic regression analysis.

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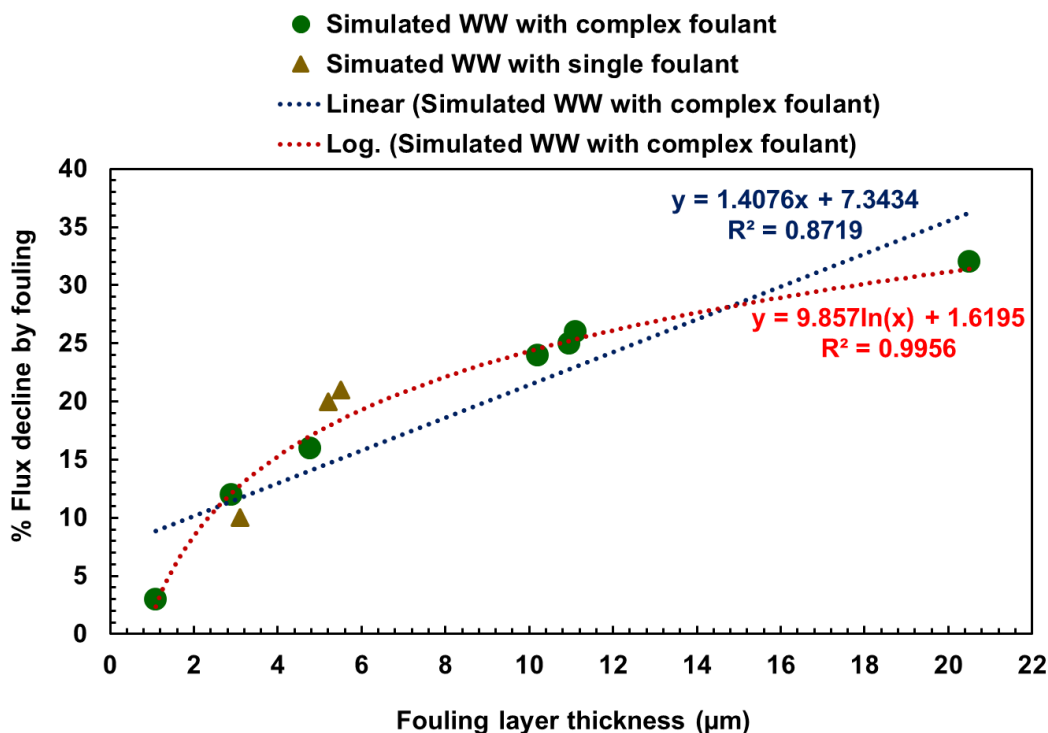
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669 **Table 4.** Pearson correlation coefficient between factors such as cake-enhanced CP, cake
 670 structural parameter, fouling resistance, fouling layer thickness and % flux decline by fouling

Factors	Cake-enhanced CP	Cake structural parameter	Fouling resistance	Fouling layer thickness	% Flux decline by fouling
Cake-enhanced CP	1	0.986**	0.873**	0.794*	0.904**
Cake structural parameter		1	0.928**	0.816**	0.925**
Fouling resistance			1	0.823**	0.908**
Fouling layer thickness				1	0.897**
% Flux decline by fouling					1

** . Correlation is significant at the 0.01 level (2-tailed) (n=10).

* . Correlation is moderate at the 0.05 level (2-tailed) (n=10).



671
 672 **Figure 9.** Correlation analysis between the fouling layer thickness and % flux decline by fouling
 673 through linear and nonlinear regressions.

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677 **4.5. Implication of this research**

678 As previously reported the raw WW/secondary WW effluents contains various types of foulants,
679 such as proteins, polysaccharides, humic substances, and lipids, with concentration of 40-140
680 mg/L [41] [42]. In this study, the presence of co-organic foulants (ALG, HA, and BSA) in the feed
681 solutions has been investigated for the FO membrane so as to give direct implications for
682 wastewater reclamation. As illustrated in Fig. 10-g the current study reveals the critical role of
683 choosing initial flux in organic fouling control and CPs minimization of FO operation. As
684 indicated by the results of Fig. 10-d, e, f, after 120 h operation the greater cake layer-ECP effect
685 was indeed observed at a high flux of 30-34 LMH whilst this fact was minor for 25 LMH. The
686 findings highlight that the classic critical (threshold) flux applies to the osmotically driven
687 membrane process. Herein our results for the TFC membrane suggest a threshold value of 25 LMH
688 for a typical WW with the organic concentration of 160 mg/L.

689 In order to develop a guide for practical operation, a summary of studies pertaining to organic
690 fouling of FO membranes is presented in Table 5. Such information provides a basis for the
691 discussion and comparison with our outcomes. For a high foulant concentration (500 mg/L), severe
692 fouling by single ALG and a low flux recovery was reported for a lab-scale TFC membrane
693 operated under a high flux of 35.2 LMH for 100 h operation [50]. Furthermore, fouling became
694 serious when operating with a simulated complex WW i.e., 400 mg/L concentration and at a
695 modest flux of 16.0 LMH. This fact is attributed to foulant-foulant-membrane interactions;
696 accelerated flux decline (43%) was observed [51]. In contrast our results indicated that a flux of
697 25 LMH is favorable when seeking to control fouling and minimize the effect of cake-ECP; this
698 outcome was with a typical WW concentration of 160 mg/L. Earlier works with laboratory-scale
699 systems are in support of this observation. Previously successful operation has been achieved with
700 a TFC membrane and a real municipal WW with flux values of 5.0 LMH [52], 10.0 LMH [53],
701 20.1 LMH [54], and 23.1 LMH [7]. However, the guideline value of 25 LMH might be
702 inappropriate for a WWs with high concentration (400-500 mg/L). Although such concentrations

703 are not to be found in a real municipal WW, high strength WW is found elsewhere e.g. dairy WW
704 or leachate WW.

705 The results summarized from Table 5 highlighted that the flux recovery by physical cleaning is
706 strongly dependent on both the initial flux level and feed composition. For example, a real dairy
707 WW containing high protein fraction posed a significant fouling i.e., flux decline by 52% and low
708 flux restoration of 81%, even with a relatively low flux of 12.2 LMH [55]. Overall findings
709 reinforce our observation that a proper choice of initial flux is essential for fouling control. To save
710 on energy, cleaning, and down-time of the installation, our present work suggests that 25 LMH is
711 a guideline for fouling control and cake-ECP minimization when operating under typical
712 conditions characterized by TFC membrane, foulant concentration of 160 mg/L and municipal
713 WW. Once high strength WW i.e., leachate WW, dairy WW was used for feed solution, it is of
714 paramount importance to re-check an appropriate initial flux for FO fouling control and cake-ECP
715 minimization.

716 In practice, a flux determined found from a laboratory flat sheet unit should only be used as a guide
717 because spiral wound modules themselves may well not achieve the same level of mass transfer
718 [38] and we note that with a pilot-scale system, fouling control was achieved with a TFC spiral
719 wound membrane as operated at a flux of 16.6 LMH [56].

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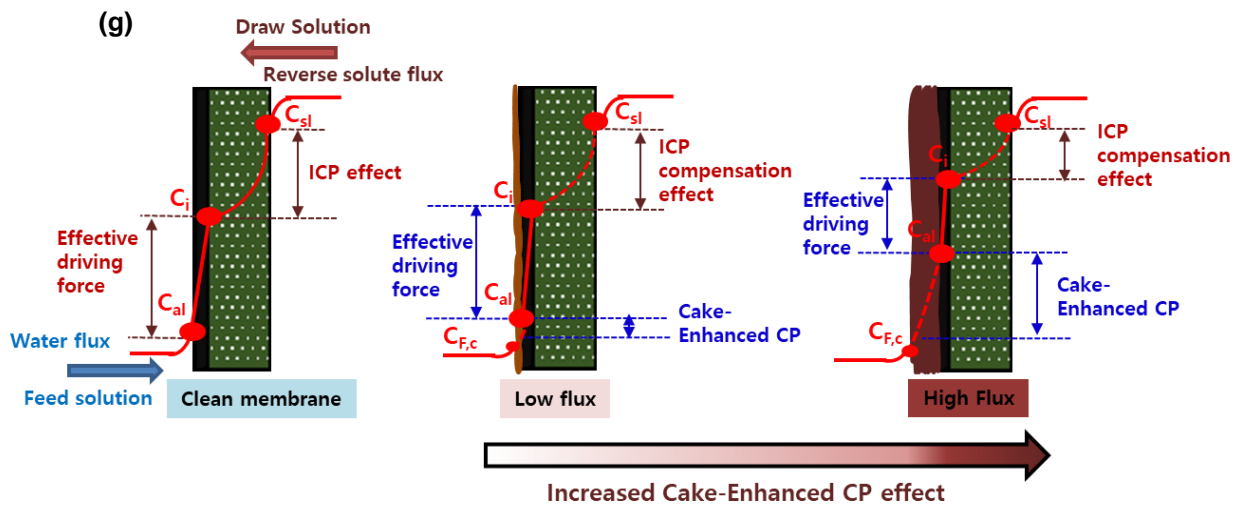
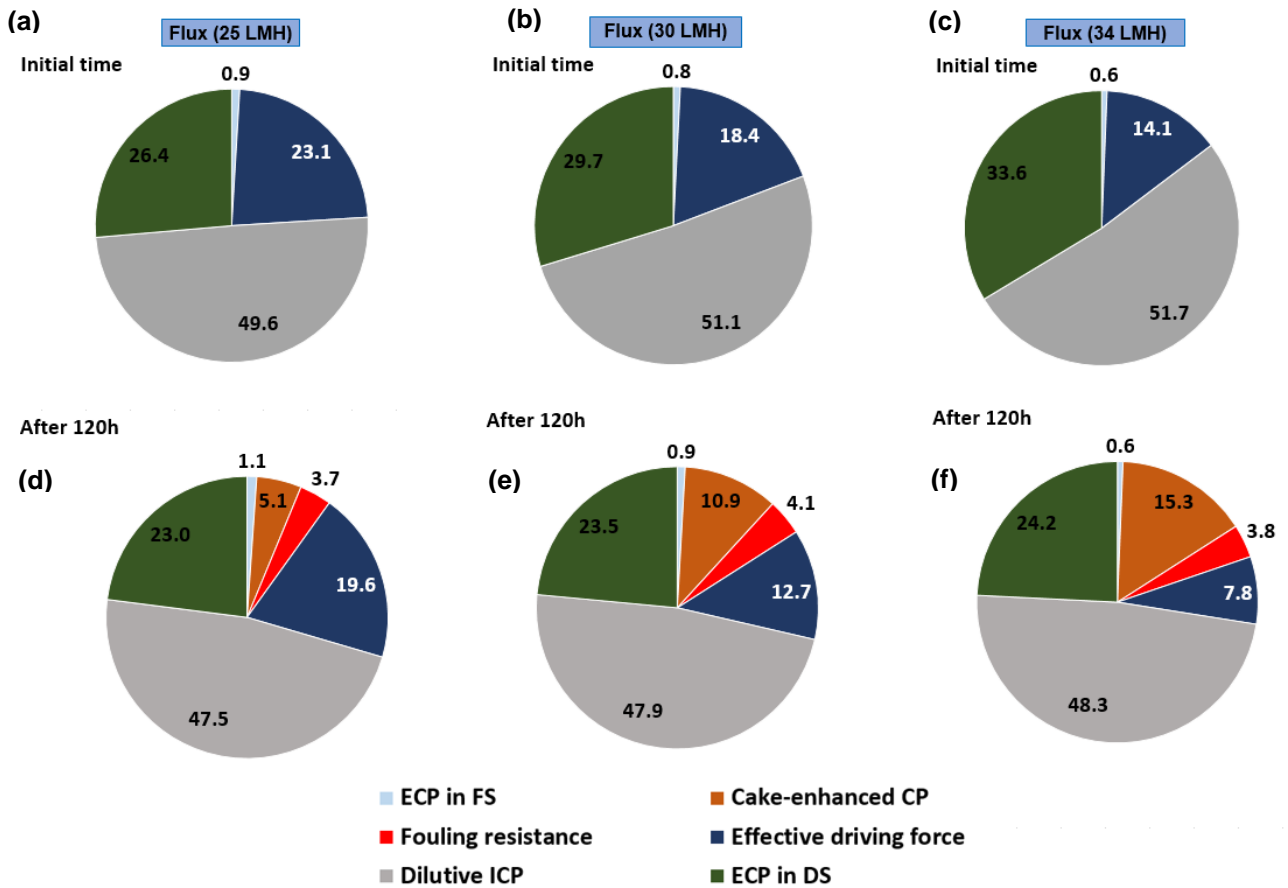
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731 **Figure 10.** Contribution of the different concentration polarizations and fouling resistance on the
 732 overall driving force under different fluxes after 120 h operation: 34 LMH (a, d) either physical or
 733 chemical cleaning was conducted after each 10 h, 30 LMH (b, e) physical cleaning was conducted
 734 after 20h, 25 LMH (c, f) physical cleaning was conducted after 40 h. Diagram illustrating the FO
 735 membrane fouled by wastewater in AL-FS orientation under different fluxes (g).

736 **Table 5.** Summary of studies on organic fouling in FO membrane

Wastewater (WW)	Draw solute	Operating conditions	Initial Water flux (LMH)	Flux decline by fouling, %	Flux recovery (%)	Reference
Simulated WW with ALG 500 g/L	NaCl= 1 M	Lab-scale TFC membrane CFV= 9.6 cm/s Operating time= 100 h	35.2	25	84	[50]
Simulated WW with 200 mg/L ALG, 200 mg/L HA and 220 mg/L CaCl ₂	Sea salt (35/L)	Lab-scale TFC membrane CFV= 10 cm/s	8.5-16	0-43	100	[51]
Secondary WW effluent	NaCl= 4%	Lab-scale TFC membrane No pretreatment CFV= 4.17 cm/s Operating time= 16 days	5.0	0	-	[52]
Municipal WW	NaCl= 5 M	Lab-scale TFC membrane Pretreatment: Coagulation, flocculation, microwaving, MF CFV= 4.17 cm/s Operating time= 4.5 h	10.0	2	-	[53]
Dairy WW	NaCl= 1 M	Lab-scale TFC membrane No pretreatment Operating time= 18 h	12.2	52	81	[55]
Municipal WW	Synthetic brine seawater	Lab-scale TFC membrane Pretreatment= MF CFV= 6.0 cm/s Operating time= 24 h	20.1	-	98	[54]
Secondary municipal WW effluent	Real seawater	Lab-scale TFC CFV= 8.5 cm/s Operating time= 80 h	23.1	5%	-	[7]
Simulated WW with 1.2 g/L sea salt, 0.22 g/L CaCl ₂ , 200 mg/L HA and 200 mg/L ALG	Sea salt (35g/L)	Pilot-scale with TFC membrane module Operating time: 3 cycles (80% recovery feed water for each cycle)	16.6	5%	-	[56]

737 Note: Flux recovery was by physical cleaning. “-” indicates absence of relevant information

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741 **5. Conclusions**

742 The current work not only explored the flux decline patterns but also quantified the synergistic
743 effect of CPs and fouling for different operating fluxes and feed compositions. The flux decline
744 pattern was dependent on the chosen operational flux, WW composition. Notably, once a chemical
745 cleaning was imposed, this induced a change of flux decline pattern. FO fouling was severe and
746 more irreversible at a high flux of 34 LMH, with the formation of a cohesive and compact cake
747 layer. Compared with operation at 25 LMH, operation at this condition triggered a four-fold
748 increase in cake structural parameter and ten-fold increase in cake layer-enhanced CP after 40 h
749 operation. This study confirmed that high protein fraction present in WW promoted FO fouling
750 and caused more irreversible fouling. Based on quantitative analysis of CPs and fouling resistance,
751 the ICP contributed most but cake layer-enhanced CP and fouling resistance had significant impact
752 at the severe fouling condition where they were found to account for 24.5% of the overall driving
753 force.

754
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761
762 **Appendix A: Supplementary data**

763
764 **Abbreviations**

765 ΔP_c : Pressure drop caused by cake layer

766 μ : Water dynamic viscosity

767 A: water permeability coefficient

768 AL-FS: Active layer-facing feed solution

769 B: solute permeability coefficient

770 C_{a1} : solute concentration at interface between foulant cake layer and the active layer surface

771 $C_{D,b}$: solute concentration of the bulk draw solution

772 C-ECP: cake layer-enhanced concentration polarization

773 CECP: concentrative external concentration polarization

774 $C_{F,b}$: solute concentration of the bulk feed solution

775 $C_{F,c}$: solute concentration on the surface of foulant cake layer

776 C_i : solute concentration at the interface between the active layer and the support layer

777 C_{s1} : solute concentration on the membrane surface of the support layer

778 D_c : diffusion coefficient in the cake layer

779 D_D : bulk diffusion coefficient of draw solution

780 DECP: dilutive external concentration polarization

781 D_F : bulk diffusion coefficient of feed solution

782 D_h : foulant pore hydraulic diameter

783 D_s : diffusion coefficient in the support layer

784 ECP: External concentration polarization

785 ICP: internal concentration polarization

786 J_w : water flux

787 k_D : mass transfer coefficient in draw side

788 k_F : mass transfer coefficient in feed side

789 RSF: reverse solute flux

790 S_c : foulant cake structural parameter

791 S_s : structural parameter of the support layer

792 δ_c : cake layer thickness

793 δ_D : thickness of the dilutive external concentration polarization in draw side

794 δ_F : thickness of the concentrative external concentration polarization in feed side

795 ε_c : cake layer porosity

796 ρ : Water density

797 τ_c : cake layer tortuosity

798

799 **References**

- 800 [1] M. Elimelech, W. a Phillip, The future of seawater desalination: energy, technology, and
801 the environment., *Science*. (2011). doi:10.1126/science.1200488.
- 802 [2] R. Valladares Linares, Z. Li, V. Yangali-Quintanilla, N. Ghaffour, G. Amy, T. Leiknes,
803 J.S. Vrouwenvelder, Life cycle cost of a hybrid forward osmosis - low pressure reverse
804 osmosis system for seawater desalination and wastewater recovery, *Water Res.* 88 (2016)
805 225–234. doi:10.1016/j.watres.2015.10.017.
- 806 [3] Q. Li, Z. Xu, I. Pinnau, Fouling of reverse osmosis membranes by biopolymers in
807 wastewater secondary effluent: Role of membrane surface properties and initial permeate
808 flux, *J. Memb. Sci.* (2007). doi:10.1016/j.memsci.2006.12.027.
- 809 [4] T.Y. Cath, N.T. Hancock, C.D. Lundin, C. Hoppe-Jones, J.E. Drewes, A multi-barrier
810 osmotic dilution process for simultaneous desalination and purification of impaired water,
811 *J. Memb. Sci.* (2010). doi:10.1016/j.memsci.2010.06.056.
- 812 [5] M. Xie, L.D. Nghiem, W.E. Price, M. Elimelech, Impact of humic acid fouling on
813 membrane performance and transport of pharmaceutically active compounds in forward
814 osmosis, *Water Res.* 47 (2013) 4567–4575. doi:10.1016/j.watres.2013.05.013.
- 815 [6] Y. Sun, J. Tian, Z. Zhao, W. Shi, D. Liu, F. Cui, Membrane fouling of forward osmosis
816 (FO) membrane for municipal wastewater treatment: A comparison between direct FO
817 and OMBR, *Water Res.* 104 (2016) 330–339. doi:10.1016/j.watres.2016.08.039.
- 818 [7] F. Volpin, E. Fons, L. Chekli, J.E. Kim, A. Jang, H.K. Shon, Hybrid forward osmosis-
819 reverse osmosis for wastewater reuse and seawater desalination: Understanding the
820 optimal feed solution to minimise fouling, *Process Saf. Environ. Prot.* 117 (2018) 523–
821 532. doi:10.1016/j.psep.2018.05.006.
- 822 [8] S.M. Iskander, S. Zou, B. Brazil, J.T. Novak, Z. He, Energy consumption by forward
823 osmosis treatment of landfill leachate for water recovery, *Waste Manag.* 63 (2017) 284–
824 291. doi:10.1016/j.wasman.2017.03.026.
- 825 [9] B. Aftab, Y.S. Ok, J. Cho, J. Hur, Targeted removal of organic foulants in landfill leachate
826 in forward osmosis system integrated with biochar/activated carbon treatment, *Water Res.*
827 160 (2019) 217–227. doi:10.1016/j.watres.2019.05.076.
- 828 [10] S. Kook, C. Lee, T.T. Nguyen, J. Lee, H.K. Shon, I.S. Kim, Serially connected forward
829 osmosis membrane elements of pressure-assisted forward osmosis-reverse osmosis hybrid
830 system: Process performance and economic analysis, *Desalination*. (2018).
831 doi:10.1016/j.desal.2018.09.019.
- 832 [11] C. Boo, M. Elimelech, S. Hong, Fouling control in a forward osmosis process integrating
833 seawater desalination and wastewater reclamation, *J. Memb. Sci.* (2013).
834 doi:10.1016/j.memsci.2013.05.004.
- 835 [12] C. Lee, N. Thanh, R. Syamsul, I.S. Kim, Performance analysis of serially-connected
836 membrane element for pressure- assisted forward osmosis : wastewater reuse and seawater
837 desalination, *Desalin. Water Treat.* 183 (2020) 104–113. doi:10.5004/dwt.2020.25260.
- 838 [13] C. Lee, T.-T. Nguyen, R.S. Adha, H.K. Shon, I.S. Kim, Influence of hydrodynamic
839 operating conditions on organic fouling of spiral-wound forward osmosis membranes:
840 Fouling-induced performance deterioration in FO-RO hybrid system, *Water Res.* 185
841 (2020) 116154. doi:10.1016/j.watres.2020.116154.
- 842 [14] M.M. Motsa, B.B. Mamba, A. D’Haese, E.M.V. Hoek, A.R.D. Verliefde, Organic fouling
843 in forward osmosis membranes: The role of feed solution chemistry and membrane
844 structural properties, *J. Memb. Sci.* 460 (2014) 99–109.
845 doi:10.1016/j.memsci.2014.02.035.
- 846 [15] M.M. Motsa, B.B. Mamba, J.M. Thwala, A.R.D. Verliefde, Osmotic backwash of fouled
847 FO membranes: Cleaning mechanisms and membrane surface properties after cleaning,
848 *Desalination*. 402 (2017) 62–71. doi:10.1016/j.desal.2016.09.018.
- 849 [16] T. Nguyen, S. Kook, C. Lee, R.W. Field, I.S. Kim, Critical flux-based membrane fouling
850 control of forward osmosis : Behavior , sustainability , and reversibility, *J. Memb. Sci.*

- 851 570–571 (2019) 380–393. doi:10.1016/j.memsci.2018.10.062.
- 852 [17] B. Mi, M. Elimelech, Chemical and physical aspects of organic fouling of forward
853 osmosis membranes, *J. Memb. Sci.* 320 (2008) 292–302.
854 doi:10.1016/j.memsci.2008.04.036.
- 855 [18] P. Zhao, B. Gao, Q. Yue, H.K. Shon, Q. Li, Fouling of forward osmosis membrane by
856 protein (BSA): effects of pH, calcium, ionic strength, initial permeate flux, membrane
857 orientation and foulant composition, *Desalin. Water Treat.* 57 (2016) 13415–13424.
858 doi:10.1080/19443994.2015.1060539.
- 859 [19] A.W. Zularisam, A.F. Ismail, M.R. Salim, M. Sakinah, H. Ozaki, The effects of natural
860 organic matter (NOM) fractions on fouling characteristics and flux recovery of
861 ultrafiltration membranes, *Desalination.* (2007). doi:10.1016/j.desal.2006.10.010.
- 862 [20] B.G. Choi, D.I. Kim, S. Hong, Fouling evaluation and mechanisms in a FO-RO hybrid
863 process for direct potable reuse, *J. Memb. Sci.* 520 (2016) 89–98.
864 doi:10.1016/j.memsci.2016.07.035.
- 865 [21] M. Zhan, G. Gwak, D.I. Kim, K. Park, S. Hong, Quantitative analysis of the irreversible
866 membrane fouling of forward osmosis during wastewater reclamation: Correlation with
867 the modified fouling index, *J. Memb. Sci.* 597 (2020) 117757.
868 doi:10.1016/j.memsci.2019.117757.
- 869 [22] M. Xie, L.D. Nghiem, W.E. Price, M. Elimelech, Impact of organic and colloidal fouling
870 on trace organic contaminant rejection by forward osmosis: Role of initial permeate flux,
871 *Desalination.* 336 (2014) 146–152. doi:10.1016/j.desal.2013.12.037.
- 872 [23] T.-T. Nguyen, C. Lee, R.W. Field, I.S. Kim, Insight into organic fouling behavior in
873 polyamide thin-film composite forward osmosis membrane: critical flux and its impact on
874 the economics of water reclamation, *J. Memb. Sci.* 606 (2020) 118118.
875 doi:10.1016/j.memsci.2020.118118.
- 876 [24] F.A. Siddiqui, Q. She, A.G. Fane, R.W. Field, Exploring the differences between forward
877 osmosis and reverse osmosis fouling, *J. Memb. Sci.* 565 (2018) 241–253.
878 doi:10.1016/j.memsci.2018.08.034.
- 879 [25] J.Y. Xiong, Z.L. Cheng, C.F. Wan, S.C. Chen, T.S. Chung, Analysis of flux reduction
880 behaviors of PRO hollow fiber membranes: Experiments, mechanisms, and implications,
881 *J. Memb. Sci.* (2016). doi:10.1016/j.memsci.2016.01.012.
- 882 [26] C.Y. Tang, Q. She, W.C.L. Lay, R. Wang, A.G. Fane, Coupled effects of internal
883 concentration polarization and fouling on flux behavior of forward osmosis membranes
884 during humic acid filtration, *J. Memb. Sci.* 354 (2010) 123–133.
885 doi:10.1016/j.memsci.2010.02.059.
- 886 [27] Y. Wang, M. Zhang, Y. Liu, Q. Xiao, S. Xu, Quantitative evaluation of concentration
887 polarization under different operating conditions for forward osmosis process, *DES.* 398
888 (2016) 106–113. doi:10.1016/j.desal.2016.07.015.
- 889 [28] C. Suh, S. Lee, Modeling reverse draw solute flux in forward osmosis with external
890 concentration polarization in both sides of the draw and feed solution, *J. Memb. Sci.* 427
891 (2013) 365–374. doi:10.1016/j.memsci.2012.08.033.
- 892 [29] E. Arkhangelsky, F. Wicaksana, C. Tang, A.A. Al-Rabiah, S.M. Al-Zahrani, R. Wang,
893 Combined organic-inorganic fouling of forward osmosis hollow fiber membranes, *Water*
894 *Res.* 46 (2012) 6329–6338. doi:10.1016/j.watres.2012.09.003.
- 895 [30] X. Xiang, S. Zou, Z. He, Energy consumption of water recovery from wastewater in a
896 submerged forward osmosis system using commercial liquid fertilizer as a draw solute,
897 *Sep. Purif. Technol.* 174 (2017) 432–438. doi:10.1016/j.seppur.2016.10.052.
- 898 [31] E.M.V. Hoek, M. Elimelech, Cake-Enhanced Concentration Polarization: A New Fouling
899 Mechanism for Salt-Rejecting Membranes, *Environ. Sci. Technol.* 37 (2003) 5581–5588.
900 doi:10.1021/es0262636.
- 901 [32] L. Zheng, W.E. Price, L.D. Nghiem, Effects of fouling on separation performance by
902 forward osmosis: the role of specific organic foulants, *Environ. Sci. Pollut. Res.* (2018) 1–

- 903 12. doi:10.1007/s11356-018-2277-7.
- 904 [33] E. Nagy, I. Hegedüs, E.W. Tow, J.H. Lienhard V, Effect of fouling on performance of
 905 pressure retarded osmosis (PRO) and forward osmosis (FO), *J. Memb. Sci.* 565 (2018)
 906 450–462. doi:10.1016/j.memsci.2018.08.039.
- 907 [34] E.W. Tow, J.H. Lienhard, Quantifying osmotic membrane fouling to enable comparisons
 908 across diverse processes, *J. Memb. Sci.* 511 (2016) 92–107.
 909 doi:10.1016/j.memsci.2016.03.040.
- 910 [35] W.A. Phillip, J.S. Yong, M. Elimelech, Reverse draw solute permeation in forward
 911 osmosis: Modeling and experiments, *Environ. Sci. Technol.* 44 (2010) 5170–5176.
 912 doi:10.1021/es100901n.
- 913 [36] R.W. Field, J.J. Wu, Mass transfer limitations in forward osmosis: Are some potential
 914 applications overhyped?, *Desalination.* 318 (2013) 118–124.
 915 doi:10.1016/j.desal.2013.01.025.
- 916 [37] N.N. Bui, J.T. Arena, J.R. McCutcheon, Proper accounting of mass transfer resistances in
 917 forward osmosis: Improving the accuracy of model predictions of structural parameter, *J.*
 918 *Memb. Sci.* 492 (2015) 289–302. doi:10.1016/j.memsci.2015.02.001.
- 919 [38] R.W. Field, F.A. Siddiqui, P. Ang, J.J. Wu, Analysis of the influence of module
 920 construction upon forward osmosis performance, *Desalination.* 431 (2017) 151–156.
 921 doi:10.1016/j.desal.2017.09.003.
- 922 [39] E.W. Tow, M.M. Rencken, J.H. Lienhard, In situ visualization of organic fouling and
 923 cleaning mechanisms in reverse osmosis and forward osmosis, *Desalination.* 399 (2016)
 924 138–147. doi:10.1016/j.desal.2016.08.024.
- 925 [40] A. Tiraferri, N.Y. Yip, A.P. Straub, S.R. Castrillon, A method for the simultaneous
 926 determination of transport and structural parameters of forward osmosis membranes, *J.*
 927 *Memb. Sci.* 444 (2013) 523–538. doi:10.1016/j.memsci.2013.05.023.
- 928 [41] M. Wang, Y. Chen, Generation and characterization of DOM in wastewater treatment
 929 processes, *Chemosphere.* 201 (2018) 96–109. doi:10.1016/j.chemosphere.2018.02.124.
- 930 [42] M. Huang, Y. Li, G. Gu, Chemical composition of organic matters in domestic
 931 wastewater, *DES.* 262 (2010) 36–42. doi:10.1016/j.desal.2010.05.037.
- 932 [43] N. Vakondios, E.E. Koukouraki, E. Diamadopoulos, Effluent organic matter (EfOM)
 933 characterization by simultaneous measurement of proteins and humic matter, *Water Res.*
 934 63 (2014) 62–70. doi:10.1016/j.watres.2014.06.011.
- 935 [44] Q. She, X. Jin, Q. Li, C.Y. Tang, Relating reverse and forward solute diffusion to
 936 membrane fouling in osmotically driven membrane processes, *Water Res.* 46 (2012)
 937 2478–2486. doi:10.1016/j.watres.2012.02.024.
- 938 [45] R.W. Field, J.J. Wu, On boundary layers and the attenuation of driving forces in forward
 939 osmosis and other membrane processes, *Desalination.* 429 (2018) 167–174.
 940 doi:10.1016/j.desal.2017.12.001.
- 941 [46] Z. Wang, J. Tang, C. Zhu, Y. Dong, Q. Wang, Z. Wu, Chemical cleaning protocols for
 942 thin film composite (TFC) polyamide forward osmosis membranes used for municipal
 943 wastewater treatment, *J. Memb. Sci.* 475 (2015) 184–192.
 944 doi:10.1016/j.memsci.2014.10.032.
- 945 [47] A. Simon, W.E. Price, L.D. Nghiem, Influence of formulated chemical cleaning reagents
 946 on the surface properties and separation efficiency of nanofiltration membranes, *J. Memb.*
 947 *Sci.* (2013). doi:10.1016/j.memsci.2012.12.029.
- 948 [48] R.W. Field, G.K. Pearce, Critical, sustainable and threshold fluxes for membrane filtration
 949 with water industry applications, *Adv. Colloid Interface Sci.* 164 (2011) 38–44.
 950 doi:10.1016/j.cis.2010.12.008.
- 951 [49] Q. She, R. Wang, A.G. Fane, C.Y. Tang, Membrane fouling in osmotically driven
 952 membrane processes: A review, *J. Memb. Sci.* 499 (2016) 201–233.
 953 doi:10.1016/j.memsci.2015.10.040.
- 954 [50] J. Jung, J. Ryu, Y. Yu, J. Kweon, Characteristics of organic fouling, reversibility by

- 955 physical cleaning and concentrates in forward osmosis membrane processes for
956 wastewater reclamation, *Chemosphere*. 245 (2020) 125787.
957 doi:10.1016/j.chemosphere.2019.125787.
- 958 [51] G. Blandin, H. Vervoort, P. Le-Clech, A.R.D. Verliefde, Fouling and cleaning of high
959 permeability forward osmosis membranes, *J. Water Process Eng.* 9 (2016) 161–169.
960 doi:10.1016/j.jwpe.2015.12.007.
- 961 [52] Z. Li, R. Valladares Linares, S. Bucs, L. Fortunato, C. Hélix-Nielsen, J.S. Vrouwenvelder,
962 N. Ghaffour, T.O. Leiknes, G. Amy, Aquaporin based biomimetic membrane in forward
963 osmosis: Chemical cleaning resistance and practical operation, *Desalination*. 420 (2017)
964 208–215. doi:10.1016/j.desal.2017.07.015.
- 965 [53] T. Hey, N. Bajraktari, J. Vogel, C. Hélix Nielsen, J. la Cour Jansen, K. Jönsson, The
966 effects of physicochemical wastewater treatment operations on forward osmosis, *Environ.*
967 *Technol. (United Kingdom)*. 38 (2017) 2130–2142. doi:10.1080/09593330.2016.1246616.
- 968 [54] S. Yang, B. Gao, A. Jang, H. kyong Shon, Q. Yue, Municipal wastewater treatment by
969 forward osmosis using seawater concentrate as draw solution, *Chemosphere*. 237 (2019)
970 124485. doi:10.1016/j.chemosphere.2019.124485.
- 971 [55] H. Song, F. Xie, W. Chen, J. Liu, FO/MD hybrid system for real dairy wastewater
972 recycling, *Environ. Technol. (United Kingdom)*. 39 (2018) 2411–2421.
973 doi:10.1080/09593330.2017.1377771.
- 974 [56] J. Kim, G. Blandin, S. Phuntsho, A. Verliefde, P. Le-clech, H. Shon, Practical
975 considerations for operability of an 8 " spiral wound forward osmosis module :
976 Hydrodynamics , fouling behaviour and cleaning strategy, *DES*. 404 (2017) 249–258.
977 doi:10.1016/j.desal.2016.11.004.
978