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1 **Mischaracterising wildlife trade and its impacts may mislead policy processes**

2
3 Daniel W.S. CHALLENGER^{1*}, Dan BROCKINGTON², Amy HINSLEY¹, Michael
4 HOFFMANN³, Jonathan E. KOLBY^{4,5}, Francis MASSÉ⁶, Daniel J.D. NATUSCH⁷,
5 Thomasina E.E. OLDFIELD⁸, Willow OUTHWAITE⁸, Michael 'T SAS-ROLFES⁹, and E.J.
6 MILNER-GULLAND¹.

7
8 **Corresponding author*

9 ¹Department of Zoology, University of Oxford, Zoology Research and Administration
10 Building, 11a Mansfield Road, Oxford, OX1 3SZ, United Kingdom
11 (*dan.challender@zoo.ox.ac.uk; +447745547585; ORCID ID: 0000-0002-0606-1715;
12 ej.milner-gulland@zoo.ox.ac.uk; amy.hinsley@zoo.ox.ac.uk)

13 ²Sheffield Institute for International Development, University of Sheffield, Sheffield, S10
14 2TN, United Kingdom (d.brockington@sheffield.ac.uk)

15 ³Conservation and Policy, Zoological Society of London, Regent's Park, London, NW1 4RY,
16 United Kingdom (mike.hoffmann@zsl.org)

17 ⁴IUCN SSC Amphibian Specialist Group, Toronto, Canada (jonathankolby@gmail.com)

18 ⁵College of Public Health, Medical and Veterinary Sciences, James Cook University,
19 Townsville, Australia.

20 ⁶Department of Geography & Environmental Sciences, Northumbria University, United
21 Kingdom (francis.masse@northumbria.ac.uk)

22 ⁷Department of Biological Sciences, Macquarie University, NSW 2109, Australia
23 (d.natusch@epicbiodiversity.com)

24 ⁸TRAFFIC, The David Attenborough Building, Pembroke Street, Cambridge CB2 3QZ,
25 United Kingdom (thomasina.oldfield@traffic.org; willow.outhwaite@traffic.org)

26 ⁹School of Geography and the Environment, University of Oxford, Oxford OX1 3QY,
27 United Kingdom (tsas.rolfes@gmail.com)

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45 **Abstract**

46 Overexploitation is a key driver of biodiversity loss but the relationship between the use and
47 trade of species and conservation outcomes is not always straightforward. Accurately
48 characterising wildlife trade and understanding the impact it has on wildlife populations are
49 therefore critical to evaluating the potential threat trade poses to species and informing local
50 to international policy responses. However, a review of recent research that uses wildlife and
51 trade-related databases to investigate these topics highlights three relatively widespread
52 issues: (1) mischaracterisation of the threat that trade poses to certain species or groups, (2)
53 misinterpretation of wildlife trade data (and illegal trade data in particular), resulting in the
54 mischaracterisation of trade, and (3) misrepresentation of international policy processes and
55 instruments. This is concerning because these studies may unwittingly misinform
56 policymaking to the detriment of conservation, for example by undermining positive
57 outcomes for species and people along wildlife supply chains. Moreover, these issues
58 demonstrate flaws in the peer-review process. As wildlife trade articles published in peer-
59 reviewed journals can be highly influential, we propose ways for authors, journal editors,
60 database managers, and policymakers to identify, understand and avoid these issues as we all
61 work towards more sustainable futures.

62

63 **Introduction**

64 Overexploitation is a key driver of biodiversity loss (Maxwell et al., 2016) but the
65 relationship between the use and trade of species and conservation outcomes is not always
66 straightforward. While harvest and trade can sometimes benefit both wildlife populations and
67 people, at other times it can drive biodiversity loss (Hutton & Leader-Williams, 2003;
68 Cooney et al., 2015). Accurate characterisation of wildlife trade and an understanding of the
69 impact it has on wildlife populations are therefore critical to evaluating the potential threat

70 trade poses to species and to informing local to international policy responses. Large-scale
71 databases are increasingly being used as tools to guide international conservation policy.
72 These include the IUCN (International Union for Conservation of Nature) Red List of
73 Threatened Species (hereafter ‘Red List’) and the CITES (Convention on International Trade
74 in Endangered Species of Wild Fauna and Flora) Trade Database. Similar databases focus on
75 illegal wildlife trade (Supplementary Material 1). There is also an increasing body of research
76 using datasets derived from monitoring of wildlife trade that takes place online (e.g., Hinsley
77 et al., 2016). Studies using these data sources frequently offer policy recommendations to
78 inform international policymaking (e.g., CITES, 2019a).

79

80 Although many examples of wildlife trade research using these datasets appropriately exist
81 (e.g., Gale et al., 2019), research studies that describe wildlife trade and its impacts – and
82 suggest policy interventions – sometimes misunderstand or misinterpret the datasets used
83 and/or inappropriately interpret the results. As research has the potential to influence
84 policymakers taking critical decisions on the sustainability and regulation of wildlife trade,
85 this is problematic. These studies also demonstrate flaws in the peer-review process, and the
86 problem is compounded when subsequent authors apply the same methodologies and make
87 identical errors.

88

89 Here, we examine a non-random selection of recent (last ~5 years) research studies and
90 discuss three key issues in wildlife trade research. These are (1) mischaracterisation of the
91 threat wildlife trade poses to species, (2) misinterpretation of wildlife trade data, and illegal
92 trade data in particular, and (3) misrepresentation of international policy processes and
93 instruments. We highlight a generic challenge for researchers through to end-users, and
94 propose ways authors, journal editors, database managers, and policymakers can identify and

95 address these issues. Our purpose is not to specifically critique the authors or their work.
96 Rather, while recognizing that many of these studies present important methodological or
97 other scientific advances, we discuss these articles because they are contemporary and
98 illustrate the issues.

99

100 **Mischaracterisation of the threat wildlife trade poses to species**

101 Various studies have characterised the threat of wildlife trade to species (Table 1). However,
102 these threats can be mischaracterised where data are misinterpreted, or results are
103 overinterpreted by authors subjectively evaluating the impact of trade on wild populations
104 without supporting evidence.

105

106 Misinterpretation of datasets can arise if researchers are not aware of important limitations to
107 the datasets they are using. Regarding the Red List for instance, not all taxonomic groups
108 have been comprehensively assessed, so there are biases in species taxonomic coverage. In
109 addition, not all species have complete data on scope and severity of threats (making it
110 difficult to distinguish the relative impacts of different threats), and information on use and
111 trade of species is incomplete for many taxa because it is not required documentation for Red
112 List assessments (IUCN, 2013). Fukushima et al. (2020), for example, analysed patterns of
113 use and trade among species on the Red List without reference to inconsistencies in
114 documentation. They further estimated the proportion of ‘traded’ species among threatened
115 species in different phyla but present the results as the proportion of species ‘threatened by
116 trade’, an incorrect and highly misleading assumption. Using the Red List to determine if use
117 and/or trade is a threat to species requires interrogating the threats classification scheme
118 (particularly scheme 5 on Biological Resource Use) and paying close attention to the
119 associated threat codings. Scheffers et al. (2019) constructed a list of ‘traded’ species by

120 combining data from the Red List with the species in the CITES Appendices and estimated
121 that trade affects 24% of terrestrial vertebrates globally. However, this mistakenly equates
122 *being in trade* with *risk of extinction from trade* and assumes that all species listed in the
123 CITES Appendices are in trade when they are not (Challender et al., 2019a). Many species
124 are included in CITES for precautionary purposes because they resemble traded species (i.e.,
125 are ‘lookalikes’; CITES Res. Conf. 9.24, Rev. CoP17), or as part of taxonomic groups where
126 the entire group is listed (higher-taxon listings; e.g., parrots [Psittaciformes spp.]).
127 Additionally, listing species in CITES Appendix I is intended to prevent commercial,
128 international trade (rather than indicate that a species is in trade). A more appropriate
129 analytical approach would have been to identify species *known* to be in trade using the
130 CITES trade data (for international trade in CITES-listed species) combined with data from
131 the Red List and other sources. Table 1 summarises these issues, together with the
132 consequences for the arguments made, and provides additional examples.

133

134 Wildlife trade can positively or negatively affect populations of wild species and
135 sustainability depends on appropriate governance of varying interactions between biological,
136 economic, and social factors (Hutton & Leader-Williams, 2003; Cooney et al., 2015).

137 Understanding the impact of trade-driven harvest on wild populations requires data on critical
138 population parameters, including intertemporal harvest rates and their influence on density
139 (Sutherland, 2001). However, various studies (Table 1) have bypassed such in-depth analyses
140 and used trade volumes subjectively to determine that trade is (or is likely to be) detrimental
141 to species populations and thus prescribed policy responses (e.g., include species in the
142 CITES Appendices). Auliya et al. (2016) discussed the impact of trade on particular reptile
143 taxa but concluded that trade in a broader range of species (whether legal or illegal) should,
144 by default, be considered detrimental to their survival. This is problematic because in many

145 cases whether trade-driven harvest is detrimental to populations remains an open question
146 requiring further research. While some species may be threatened by modest levels of trade,
147 others can be traded in large volumes without trade posing a threat to the survival of the
148 species in the wild (e.g., reticulated python *Malayopython reticulatus* and American Alligator
149 *Alligator mississippiensis*; Natusch et al., 2016; Joanen et al., 2021).

150

151 **Misinterpretation of wildlife trade data**

152 Since 2010, ~130 studies have used the CITES Trade Database to characterise international
153 wildlife trade (UNEP-WCMC, unpubl. data); others have used the US Fish and Wildlife
154 Service Law Enforcement Management Information System (LEMIS) data, or other
155 databases (Supplementary Material 1). However, numerous studies have misinterpreted these
156 databases, resulting in the mischaracterisation of trade dynamics and volumes (Table 1). For
157 example, a common error is treating each row of data in the ‘comparative tabulation’ output
158 from the CITES Trade Database (which may comprise many shipments aggregated into a
159 single row) as a single trade transaction, which miscalculates transaction frequency (Table 1).
160 Similar misinterpretation applies to LEMIS data, which records trade in all wildlife species
161 that cross US borders. Sosnowski and Petrossian (2020) analysed seizures of fashion-related
162 wildlife products in the US but inflated the number of seizures. They assumed each row of
163 data represented a single seizure, but whether a single seizure is represented by one or more
164 data rows varies. For example, a single confiscated item derived from more than one wildlife
165 species will appear as multiple rows of LEMIS data and should not be counted as multiple
166 seizure events (Natusch et al., 2021). Failure to correctly interpret the number of seizures
167 and/or items seized can erroneously inflate the extent of illegal trade.

168

169 Another problem regarding CITES trade data is interpretation of source code 'I', which has
170 been used to describe illegal international trade dynamics (Table 1). This code can refer to
171 seizures made due to a lack of valid permits accompanying specimens in trade, or
172 international trade in specimens of species that have previously been seized or confiscated
173 but are being legally exported in accordance with CITES Res. Conf. 17.8 paragraph 8 (e.g.,
174 repatriation to the source country). Hence, the code may or may not indicate illegal trade.
175 Without verification from the relevant CITES Management Authorities that trade records do
176 indeed refer to illegal trade it is not possible to accurately characterise illegal trade using
177 these data. Alternative illegal wildlife trade datasets exist (Supplementary Material 1).

178

179 Analysis of seizure data is frequently used to understand illegal wildlife trade, but
180 misinterpretation of these data is commonplace (Table 1). While seizure data can be useful to
181 gain insights into illegal trade dynamics, they suffer from inherent biases related to
182 enforcement effort (e.g., resources committed), rates of seizure (proportion of illegal
183 transactions seized) and reporting (proportion of seizures reported to focal database), which
184 differ between countries (Underwood et al., 2013). Critically, these biases need to be
185 appropriately accounted for in order to derive meaningful temporal trade trends or spatial
186 patterns. Underwood et al. (2013) used Bayesian hierarchical latent variable modelling to
187 account for biases and produce relative trends in illegal international trade in elephant ivory
188 using ETIS (Elephant Trade Information System) data. Similar analyses have not been
189 completed for other species, in part because of the large and comprehensive datasets needed
190 (ETIS holds >29,000 seizure records; TRAFFIC, 2019). Yet researchers commonly fail to
191 recognise (or account for) these biases explicitly and/or incorrectly describe illegal trade
192 trends from the raw data in qualitative terms (e.g., illegal trade is increasing) without the
193 necessary caveats. These 'trends' are not meaningful. For example, an apparent increase in

194 seizures may reflect greater law enforcement effort or discovery of a previously unknown
195 smuggling method rather than an increase in illegal trade.

196

197 Seizure data can be used to: i) estimate the minimum number of individual animals or plants
198 in illegal trade, ii) estimate minimum volumes or quantities of derivatives over a defined
199 period, and iii) characterise spatial trafficking patterns (e.g., countries of origin, export,
200 transit, and destination) based on reported seizures. However, studies using seizure data for
201 these purposes should explicitly acknowledge the inherent biases and the fact that the data
202 reflect known seizures, rather than absolute trade volumes or bias-adjusted trends or spatial
203 patterns.

204

205 **Misrepresentation of international policy processes and instruments**

206 The framing of research can result in misguided recommendations, stemming in part from
207 authors misunderstanding international policy processes and how policy instruments function.
208 Frank and Wilcove (2019), for example, estimated that it takes ~10 years for species they
209 determined to be threatened by international trade on the Red List to be included in CITES,
210 and argued for a ‘near-automatic pathway by which unprotected species identified by the
211 IUCN as threatened by international trade receive a vote for inclusion in CITES Appendix I
212 or II’. However, this seemingly simple, but ultimately far-reaching, recommendation
213 discounts four main issues, three of which are characteristic of other studies. First, the Red
214 List and CITES apply independent (albeit related) criteria for determining threat status; the
215 Red List sets quantitative thresholds for species to be listed in a particular Red List category,
216 while the CITES listing criteria only provide indicative, non-binding guidelines on numerical
217 values (see Annex 5 of Res. Conf. 9.24, [Rev. CoP17]). Consequently, a species determined
218 to be threatened by international trade according to the Red List may, or may not, qualify for

219 inclusion in CITES (Challender et al., 2019b). Other articles have also made this assumption
220 (e.g., Gorobets, 2020). Second, Frank and Wilcove (2019) focus on Appendix I and II only,
221 overlooking Appendix III. Parties to CITES may unilaterally include species in Appendix III
222 without the lengthy process that would be required for proposing species be included in
223 Appendices I and II, which would reduce the time-lags identified by the authors.

224

225 Third, the establishment of a ‘near-automatic pathway’ would require fundamental changes to
226 the Convention, probably including amendment of the Convention text, requiring the
227 agreement of the Parties. However, the feasibility and political palatability of the proposal
228 were not considered by the authors. This is non-trivial because even suggestions agreed by
229 the Parties can take many years to take effect. The Gaborone amendment allowing regional
230 economic integration organisations to accede to CITES took 30 years to enter into force
231 following its adoption (CITES, 2013). Other studies apply a similar approach to suggested
232 reforms to wildlife trade regulation (Marshall et al., 2020), including the ‘reverse listing’
233 model (Altherr and Lameter, 2020), whereby all international trade would be prohibited
234 unless it could be demonstrated to be sustainable. Scientific research should be used to
235 inform potential wildlife trade policy reforms, but such studies should consider the realities of
236 the policy frameworks discussed.

237

238 Fourth, Frank and Wilcove (2019) suggest that including species in CITES Appendix I or II
239 may help to avoid the extinction of species, but they fail to acknowledge that such measures
240 may at times do more harm than good. Although designed to restrict trade and reduce
241 unsustainable harvesting, such listings may signal scarcity to speculative collectors,
242 stockpilers, and organised crime groups, and at least in theory could lead to scarcity-driven
243 price increases that in turn raise incentives for accelerated wild harvest (e.g., Asian arowana

244 *Scleropages formosus*; Crockett, 2021; Bergstrom, 1990; Courchamp et al., 2006). The
245 assumption that including species in CITES is positive for their conservation is common in
246 the wildlife trade literature. This includes articles which recommend that species be included
247 in the Appendices but fail to evaluate realistically whether it would be positive for those
248 species and how this may change over time (e.g., Shepherd et al., 2019; Table 1). Evaluating
249 the potential conservation benefits and risks to including species in CITES requires an in-
250 depth understanding of the social-ecological system in which harvest, trade and consumption
251 of species occur (e.g., using theories of change; Cooney et al., 2021). Future research which
252 considers CITES as a conservation tool should explicitly evaluate both the potential
253 conservation benefits and risks of including species in the Convention.

254

255 **Addressing the mischaracterisation of wildlife trade**

256 The publication and dissemination of research that mischaracterises wildlife trade and its
257 impact, and/or misrepresents policy processes and instruments is concerning for two main
258 reasons. First, this research may unwittingly misinform or misdirect wildlife trade policy and
259 associated action by government agencies and conservation practitioners (at local to
260 international scales), including the misallocation of resources. Such research may be
261 interpreted uncritically by policymakers and practitioners who may not have the time or
262 expertise to critically evaluate the methodologies used. This could lead to policy which
263 undermines positive outcomes for species and associated benefits for people along wildlife
264 supply chains, contributing to achievement of the Sustainable Development Goals (Booth et
265 al., 2021). More broadly, this research may not contribute towards improved public
266 understanding because the associated press coverage can repeat errors made in publications
267 (e.g., Dunphy, 2019).

268

269 Second, the articles discussed demonstrate certain flaws in the peer-review process.
270 Researchers may publish responses, but rebuttals seldom alter scientific or public perceptions
271 of original articles (Banobi et al., 2011), and readers of an article are rarely made aware that a
272 response has been published. Even where they do exist, responses are typically limited in
273 terms of space, especially in high-impact journals, meaning it is not always possible to fully
274 address the problems identified. Once published, the original articles continue to be cited
275 (Cosentino and Veríssimo, 2016) and influence the conservation agenda, to the potential
276 detriment of the science-policy interface.

277

278 To avoid the issues discussed in future research, we propose the following measures for
279 researchers, journal editors, database managers, and policymakers.

280

281 For researchers: Researchers should familiarise themselves with the datasets they will use
282 before starting their research, to avoid misinterpretation and so they are aware of important
283 limitations and biases. Guidance accompanies various online databases including the CITES
284 Trade Database (UNEP-WCMC, 2013; CITES, 2019b; and see Robinson and Sinovas, 2018)
285 and IUCN maintains protocols and guidance documents pertaining to Red List data (e.g.,
286 IUCN, 2013). There are also resources on interpretation of illegal trade data (e.g., TRAFFIC,
287 2019). Uncertainties concerning the extraction, download, use, and/or interpretation of such
288 datasets should be clarified with database providers and managers, and/or with other
289 academics and CITES Management Authorities (e.g., for CITES source codes).

290

291 Researchers should report limitations in the data accurately and any associated caveats, as
292 well as manipulations of the raw data they have made, when presenting analysis or
293 interpretation. Researchers should consider the biological significance of their results and

294 whether use and/or trade represents a risk for species conservation or not, or if there is
295 insufficient evidence to objectively determine the risk. Language is also important and we
296 urge care in its use. For example, a species being used for subsistence purposes does not
297 equate to a species being in trade unless it is purchased/bartered for; being in trade does not
298 mean that trade crosses international borders (though note that ‘trade’ within CITES does
299 refer to international trade); and a species in use or trade is not automatically threatened by
300 this use/trade. More evidence-based interpretation and reporting around use and trade will
301 help to ensure that policy deliberations are well-targeted and that management interventions
302 work for species conservation.

303

304 If making policy recommendations, authors should acquaint themselves with the treaties and
305 institutions involved, and with the broader policy and regulatory landscape, to avoid
306 misrepresenting policy processes and instruments. This could be achieved by dialogue with
307 experts in relevant institutions (e.g., IUCN, CITES, and UNEP-WCMC). Critically,
308 researchers should evaluate whether their recommendations (and implementation thereof)
309 would in fact contribute to the conservation of species, or not, and explicitly consider areas of
310 uncertainty and any associated risks (e.g., of CITES listings). If suggesting broader policy
311 reforms (e.g., to treaties) researchers should also offer evaluation of how realistic their
312 recommendations are; considering, for example, timelines, feasibility, and expected impact.
313 This would hopefully result in more robust and informed recommendations.

314

315 For journal editors: Journal editors can best ensure the correct use and analysis of wildlife
316 trade datasets by selecting peer-reviewers with in-depth knowledge of particular databases
317 and/or methods used, or the policy instruments involved. These could include individuals
318 with particular expertise (e.g., database managers), many of whom already sit on journal

319 editorial boards, and could therefore be consulted on appropriate uses of data and possible
320 reviewers. Conflicts of interest could be managed to ensure these individuals do not unduly
321 influence the publishing process. While we are not suggesting that the articles we use as
322 examples should be retracted, where wildlife trade articles are published in the future and
323 post-publication review highlights very serious errors in the methods or data analyses which
324 materially and fundamentally affect the key results and/or conclusions, journals could
325 consider retractions as an option, as is done in other disciplines (e.g., medicine) to prevent
326 perpetuation of the harmful errors. Responses which highlight key analytical issues should be
327 presented alongside original articles and made available under open access terms.

328

329 For database managers: To facilitate accurate and robust analysis of data on wildlife trade,
330 database managers should provide accessible, up-to-date guidance on the use and misuse of
331 the data they manage, including examples of best practice. Where feasible (e.g., subject to
332 resources) data managers and/or compilers should engage with researchers to develop
333 methodologically sound analyses and support correct interpretation of the data.

334

335 For policymakers and civil society organisations: It is important to critically evaluate research
336 before taking a position on an issue, in order to identify methodological errors, especially
337 where these may materially influence the results and conclusions. It is worth checking if any
338 responses to specific articles have been posted online that refute or invalidate the research
339 findings, or if articles have been retracted. If in doubt, and where important policy decisions
340 are being made, policymakers should seek assurances from the authors and independent
341 experts, including the managers of the datasets in question, on the validity of the results.

342

343 There is broad research interest in the use and trade of wildlife species. The intention behind
344 this article is not to discourage or criticise much-needed independent research in this field.
345 We strongly support ongoing innovative and exploratory research but emphasise the need for
346 care and caution in analysis, interpretation, and discussion of results – and in making policy
347 recommendations. Specifically, we want to highlight that using datasets in this space
348 (especially those that are publicly available) may require specialist analysis (Dobson et al.
349 2020). Researchers should be encouraged to take advantage of these datasets, but they should
350 do this with due consideration, aware of the broader policy context and of the potential
351 pitfalls of using secondary data. More effective communication between data generators,
352 analysts, and users would lead to more pertinent, more meaningful, and ultimately more
353 impactful science that is better positioned to make a positive contribution to conservation.

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363

364 **Author contributions**

365 All authors contributed to the conception, writing, editing and reviewing of the manuscript.

366

367 **Ethics statement**

368 No primary data were collected for this manuscript and an ethical review process was not
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370

371 **Data accessibility statement**

372 No primary data were collected for this manuscript.

373

374 **Conflict of interest**

375 The authors declare no conflict of interest.

376

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513 **Table 1. Issues, consequences for the arguments made by authors, and example articles**

Summary	Issue	Consequences for argument made	Examples*
Mischaracterisation of threat from trade	Assuming use/trade constitutes a threat to species or is detrimental to wild populations	Mischaracterizes species as threatened by trade or assumes that trade is detrimental to wild populations when this may not necessarily be the case	Auliya et al. (2016), Fukushima et al. (2020), Harrington et al. (2016), Jensen et al. (2019), Luiselli et al. (2012), Marshall et al. (2020), Scheffers et al. (2019)
Misinterpretation of wildlife trade data	Misinterpretation of CITES trade data:		
	<i>Incorrectly comparing importer- and exporter-reported data</i>	Assumes that differences in quantity between importer- and exporter-reported data reflect reporting issues, errors or 'missing data' when they may result from legitimate differences (e.g., differences in source or purpose code)	Andersson and Gibson (2018), Berec et al. (2018)
	<i>Incorrectly combining direct and indirect trade data</i>	May inflate trade volumes for items that have been imported and (re-)exported several times by treating each export/re-export as a new item	Poole and Shepherd (2016)
	<i>Misunderstanding the use of source and purpose codes†</i>	Mischaracterises trade (e.g., conflates source of specimens/purpose of trade)	Nellemann et al. (2018)
	<i>Conflation in use of terms/units‡ E.g., misinterpreting blank units in trade records as missing data (and assuming trade involved the recommended unit for specific derivatives e.g., kg) rather than 'number of items'</i>	Mischaracterises trade volumes (e.g., inflating the number of individual animals or plants in trade)	Andersson and Gibson (2018)
	<i>Assuming each row of data comprises a single shipment/incident‡</i>	Miscalculates transaction frequency since rows in e.g. a comparative tabulation output may contain multiple records (see Pavitt et al. 2018)	Berec et al. (2018), Can et al. (2019), D'Cruze and Macdonald (2015), Vall-Llosera and Su (2018)
	<i>Assuming source code I refers to illegal trade</i>	Misrepresents illegal trade levels (e.g., number of individual animals or plants involved)	D'Cruze and Macdonald (2015, 2016), Ribeiro et al. (2019), Ye et al. (2020)
	Misinterpretation of LEMIS data:		
	<i>Treating each row of data as a single seizure event</i>	Mistakenly inflates the number of seizures and thereby the extent of illegal trade	Govenechea and Indenbaum (2015), Petrossian et al. (2016), Petrossian et al. (2020), Sosnowski and Petrossian (2020)

Misinterpretation of seizure data:

Failing to acknowledge and/or account for inherent biases in seizure data and describing illegal trade as increasing or similar

Misrepresents illegal trade data and the ‘trends’ derived are not meaningful

C4ADS (2020), Hitchens and Blakeslee (2020), Morcatty et al. (2020), Paudel et al. (2020), Siriwat and Nijman (2018), UNODC (2020), Wildlife Justice Commission (2020)

Misrepresentation of international policy processes and instruments

Assuming species determined to be threatened by international trade based on the Red List would automatically qualify for inclusion in CITES

Overlooks the fact that IUCN and CITES have independent criteria and processes for determining threat status

Frank and Wilcove (2019), Gorobets (2020)

Assuming all species included in the CITES Appendices are traded

Misrepresents the CITES Appendices and inflates the number of CITES-listed species considered to be in trade

Scheffers et al. (2019)

Failing to consider the feasibility of recommended changes to policy instruments (e.g., CITES) or trade regulations

The utility of recommendations is difficult to determine because they have not been evaluated in realistic terms

Altherr and Lameter (2020), Frank and Wilcove (2019), Marshall et al. (2020).

Assuming the inclusion of species in CITES will be positive for their conservation

This assumption is misleading because the inclusion of species in CITES Appendix I or II may be positive or negative for species and change over time

Frank and Wilcove (2019), Harrington et al. (2016), Gomez (2021), Rowley et al. (2016), Shepherd et al. (2018, 2019)

514 *See Supplementary Material 2 for full citations of articles not discussed in the main text.

515 †Based on Robinson and Sinovas (2018).

516 ‡Since 2019 shipment-level trade data (with anonymised permit numbers) has been made available as a static download from the CITES Trade
517 Database, updated once a year.

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Supplementary Material 1. Known databases focusing on illegal wildlife trade.*

Geographic focus	Database	Description	URL (where applicable)	Publicly available (Y/N)
Global	UNODC World WISE (World Wildlife Seizures) database	Global database of seizures compiled from a multitude of sources, including the CITES Trade Database, CITES Annual Illegal Trade Reports, LEMIS, EU-TWIX, ETIS, World Customs Organisation, government agencies, NGOs, and from UNODC fieldwork.	https://www.unodc.org/unodc/en/data-and-analysis/wildlife.html	N
	TRAFFIC's Wildlife Trade Portal	Global database of open-source wildlife seizure and incident data.	www.wildlifetradeportal.org	Y
	ETIS (Elephant Trade Information System)	Database on seizures of elephant specimens, i.e., including but not limited to ivory, that have been made globally since 1989. The database is supported by various secondary databases which include data on law enforcement effort and efficiency, rates of reporting, domestic ivory markets, and background economic variables.	https://cites.org/eng/prog/eti`/index.php	N
	World Customs Organization, Customs Enforcement Network Database	A database of seizures, offences, and pictures facilitating analysis of trafficking by agencies with Customs expertise.		N
	Great Apes seizure database	Database of seizures and unlawful situations involving great apes and their parts and carcasses.	https://database.un-grasp.org/	N
	Environmental Investigation Agency (EIA)	Global database of seizures involving elephants, rhinos, pangolins, and big cats.		N
	Wildlife Justice Commission	Global database of seizures on selected species (e.g., pangolins)		N
	Researcher databases†	Seizure databases built by researchers (e.g., from reports of seizures online) which focus on individual species or species groups globally.		N
Regional	EU-TWIX (European Union – Trade in Wildlife Information eXchange)	Database on seizures and offences involving wildlife reported by EU Member States.	https://www.eu-twix.org/	N

	ASEAN-WEN (Association of Southeast Asian Nations-Wildlife Enforcement Network) Seizures database	Database of seizures involving wildlife in the ASEAN region as reported to the ASEAN-WEN Secretariat by law enforcement agencies.		N
National	CITES Annual Illegal Trade Reports	Database of seizures reported by CITES Parties in their annual illegal trade reports as per CITES Res. Conf. 11.17 (Rev. CoP18) on <i>National Reports</i> . Parties are required to submit reports on seizures for violations involving CITES-listed species, including those made domestically and at international borders.	For information see: https://www.cites.org/eng/resources/reports/Annual_Illegal_trade_report	N
	US Fish and Wildlife Service Law Enforcement Management Information System (LEMIS)	Database on US imports and exports of live organisms and other wildlife products.		N (except on request via FOIA)
	Other national seizure databases focused on wildlife products held by government agencies	Databases of seizures involving wildlife products made within country.		N

*A range of other organisations likely maintain seizure databases focusing on illegal wildlife trade. See UNODC: https://www.unodc.org/documents/data-and-analysis/wildlife/2020/WWCR2_Methods_Annex.pdf.

†These databases are also known to exist with a focus on particular regions (e.g., ASEAN) or countries only.

1 **Supplementary Material 2. Full citations of articles included in Table 1 but not**
2 **discussed in the main text.**

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