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Carbon-doped flower-like Bi₂WO₆ decorated carbon nanosphere nanocomposites with enhanced visible light photocatalytic degradation of tetracycline

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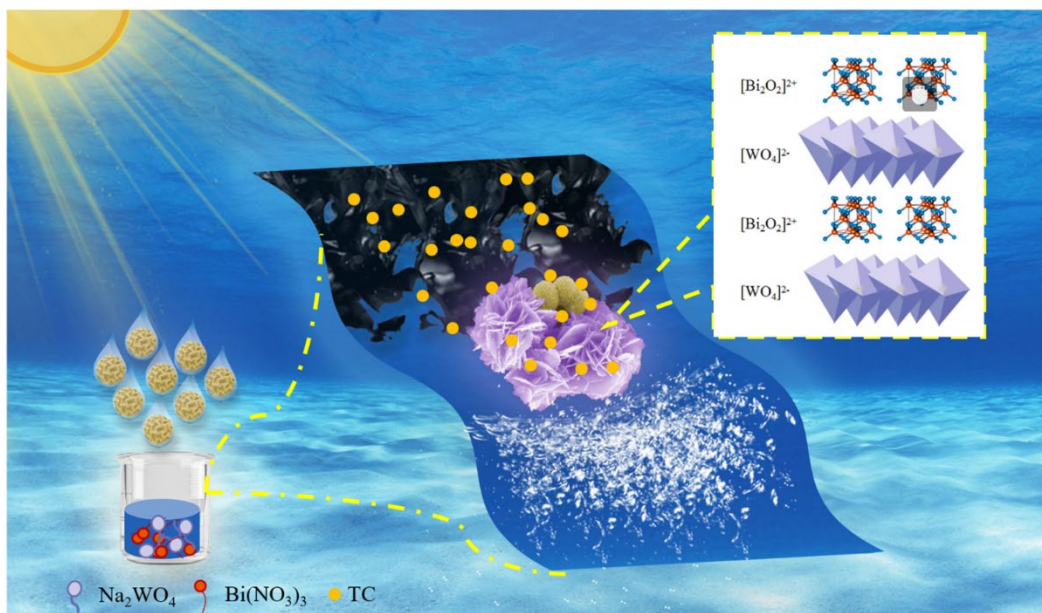
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Abstract

In search of a recyclable catalyst with synergistic adsorption and photocatalysis, unique composite photocatalysts of flower-like bismuth tungstate (Bi_2WO_6) and carbon nanospheres (CSs) were composited using a hydrothermal synthesis method (named CSs- Bi_2WO_6). Notably, based on the high visible light utilization and a reasonable band gap (2.53 eV), CSs- Bi_2WO_6 have good photocatalytic properties. For example, the composite with an optimized ratio (2% CSs- Bi_2WO_6) showed good adsorption and photocatalytic performance. Under simulated natural light conditions, the degradation rate of tetracycline (TC) by 2% CSs- Bi_2WO_6 was 84.6% in 60 min, which is nearly 25% higher than pure Bi_2WO_6 . After five cycles, the observed barely decreased TC degradation rate of 2% CSs- Bi_2WO_6 confirmed the high cyclability and reproducibility of the photocatalyst. The total organic carbon estimation of the post-degradation reaction medium corresponded to 68.2% mineralization. Furthermore, we determined the photocatalytic reaction path by LC-MS, which confirmed that the composite catalyst could effectively degrade TC molecules into small molecules. It can be concluded that the catalyst has a broad application prospect in the field of wastewater treatment.

Keyword: Flower-like Bi_2WO_6 , Carbon nanospheres, Synergistic effect, Wastewater treatment, Degradation pathways



Graphical abstract: Carbon-doped Bi_2WO_6 composite catalyst was used to degrade TC in wastewater efficiently

1. Introduction

With the acceleration of global industrialization, environmental pollution has brought serious and inevitable questions to human life [1-5]. The polluted water seeps into the ground, which further leads to different degrees of pollution of soil and groundwater [6, 7]. Therefore, environmental remediation is a necessary task for human beings to survive on Earth. In the past few decades, antibiotics have been diffusely used in the field of biological disease control. Among them, TC has high efficiency in the inhibition and even removal of a variety of pathogenic bacteria, ranking second in the list of global antibiotic production and consumption [8, 9]. Because TC has a stable aromatic ring molecular structure, it is difficult to degrade under natural conditions, which is easy to increase the resistance of receptors, and then threaten human health [10, 11]. Therefore, it is critical to develop efficient and cost-effective methods to control TC in the environment.

Among many methods of sewage treatment, adsorption is considered to be one of the key methods to alleviate water pollution due to its high efficiency and recyclability [12]. We note that the applications of carbon materials in supercapacitors, energy storage, drug delivery, and especially catalysis are widely concerned [13-15]. Since the carbon component can provide high specific surface area and a high visible light utilization, which will effectively enhance the absorption of visible light. In addition, they can also form strong interfacial electronic effects with semiconductors [16, 17], which play a positive role in improving the degradation performance of catalysts. In general, the easy stacking of most carbon materials in the preparation process is an important reason that affects the material properties. Therefore, the preparation of uniform and

stable carbon materials is our key research topic. On the other hand, carbon materials can only adsorb pollutants in the environment and cannot be mineralized into small molecules, thus producing secondary pollution, which limits their adhibition [18, 19]. To solve the problems we mentioned, a strategy of combining adsorption and photocatalysis was proposed: the pollutants are enriched on the surface of carbon materials, and it was further decomposed and mineralized by photocatalysts into small molecules [20, 21].

In recent years, research on photocatalysis technology in energy conversion and environmental remediation has been in full swing [22, 23]. Many Bi-based photocatalysts have been modified by carbonaceous materials [24, 25]. Bi_2WO_6 is one of the semiconductor materials with environmental protection, high light stability, and nontoxicity. Due to its special sandwiches-like layered structure of $[\text{Bi}_2\text{O}_2]^{2+}$ and $[\text{WO}_6]^{2-}$, Bi_2WO_6 has a better optical property and stability than monolayer catalysts [26]. The valence band of Bi_2WO_6 is formed by the hybridization of Bi 6s orbital and O 2p orbital, and the W 5d orbital also constitutes the conduction band in the meantime [27-29]. It shows a suitable band gap value (2.75 eV) and visible light response range for photocatalytic reactions [30]. It is becoming one of the most promising photocatalysts. However, as a result of the fast recombination rate of photogenerated e^- and h^+ , the application of pure Bi_2WO_6 photocatalyst is greatly limited [31]. According to many reports, we have noted that carbon materials have the function of capturing and transporting photogenerated electrons, thus improving photocatalytic performance. For instance, Li et al. [32] modified the Bi_2WO_6 material with carbon materials, and the results showed that the degradation rate of TC (20 mg/L) was 87% in 70 min. Cai et al. reported a direct S-scheme electron transfer mechanism in CZS/CDs/BWO

composite with carbon dots as the electron bridge, which exhibited a wider absorption wavelength compared to the pristine Bi_2WO_6 [33]. Similarly, other photocatalysts based on Bi_2WO_6 composite with carbon materials, such as CDs/Cl- Bi_2WO_6 [34], SCFs/BWO [35], BWO QDs/MCNOs [36], and $\text{Bi}_2\text{WO}_6/\text{rGO}$ [37] were found to exhibit photocatalytic degradation. The above studies have well proven its excellent photocatalytic activity.

In this paper, the composite photocatalyst of CSs- Bi_2WO_6 was prepared by hydrothermal synthesis. It is worth noting that the high dispersion of CSs avoids the aggregation and stacking of catalysts. Moreover, the photocatalytic performances of CSs- Bi_2WO_6 with different proportions have been measured by the photodegradation of TC solution, which presented greatly enhanced photocatalytic activity compared with pure Bi_2WO_6 . The composite catalyst has an excellent photocatalytic degradation effect. Furthermore, according to the intermediate substances produced in the degradation process, a possible photodegradation mechanism driven by visible light was proposed.

2 Experiment section

2.1 Chemicals and Characterization

Detailed information on experimental reagents and characterization was provided in the supporting information.

2.2 Preparation of CSs

CSs were prepared by conventional methods [38], 4 g of D-(+) -glucose was added to 30 mL of deionized water. After stirring for 30 min, the mixture was transferred to a 50 mL Teflon-lined steel autoclave and heated to 180 °C for 8 h. The black - brown carbonaceous material obtained by the hydrothermal reaction was washed three times by centrifugation with water and ethanol, respectively. Finally, the centrifuged products were dried in an oven at 80 °c for 6 h.

2.3 Preparation of CSs-Bi₂WO₆

Bi₂WO₆ were prepared according to the hydrothermal procedure as described elsewhere [39]. In short, about 0.5 mol of Bi(NO₃)₃ 5H₂O and 0.5 mol Na₂WO₄ 2H₂O were dissolved in 25 mL of ethylene glycol and stirred to produce a white precipitate. The CSs were dissolved in 10 mL distilled water and then sonicated for 10 min, which was then added dropwise to the above solution. The mixture was then transferred to a Teflon-lined and heated at 140 °C for 14 hours. The samples obtained were washed three times with ethanol and deionized water respectively and then centrifuged. The repeatedly washed sediment was dried at 60 °C. The composite catalysts with

different CSs contents were labeled as Bi₂WO₆, 0.5% CSs-Bi₂WO₆, 1% CSs-Bi₂WO₆, 2% CSs-Bi₂WO₆, 4% CSs-Bi₂WO₆, 8% CSs-Bi₂WO₆.

2.4 Photocatalytic TC degradation

Details of photocatalytic TC degradation are provided in the supporting information.

2.5 Adsorption kinetics study

Details of adsorption kinetics are provided in the supporting information.

3 Results and discussion

3.1 Morphology and structures

Fig. 1a shows that Bi_2WO_6 is a typical ‘flower-like’ structure with a diameter of 3-4 μm . It consists of a large number of nanosheets that cross over and come together to form a 3D structure. As we can see from Fig. 1b, CSs and Bi_2WO_6 are interconnected. However, CSs were clustered on the surface of Bi_2WO_6 , and the end is easier to be covered [40]. The primary reason for this special structure may be that the Bi_2WO_6 sheet is stacked at one end and the stress distribution on the three-dimensional structure surface is not uniform. The structure and morphology of the prepared CSs are shown in Fig. 1c, from which we can find that the CSs is a uniform and well-dispersed nanospheres with an average diameter of about 450 nm. The particle size distribution was shown in Fig. S1 in supporting information. The surface of CSs in Fig. 1b changed from smooth to rough, which may be affected by the growth of Bi_2WO_6 [41]. It could not prevent the CSs from forming close contact with Bi_2WO_6 to form a ‘flower-spheres’ structure, and will even be more conducive to the surface adsorption of organic pollutants. However, if the content of CSs increases, it is easy to form large clusters. This will affect the interaction between CSs and Bi_2WO_6 . The EDS results in Fig. 1d-1g demonstrates the distribution of C, O, W, and Bi elements in the CSs- Bi_2WO_6 composites, indicating the successful production of the composite catalysts.

In Fig. 1h, it can be seen that the 2% CSs- Bi_2WO_6 composite structure exhibits the ‘flower-spheres’ morphology. The layered structure of Bi_2WO_6 can be seen in the picture as consisting of a large number of nanosheets. The CSs are attached to the nanosheets. The TEM image of pure

Bi_2WO_6 and the magnified structure of the 2% CSs- Bi_2WO_6 in Fig. 1h were shown in supporting information (Fig. S2). Fig. 1i shows the HRTEM image of Bi_2WO_6 . We observed lattice fringes of 0.27 nm and 0.32 nm matched perfectly with the (020) and (131) crystal faces of Bi_2WO_6 , which proved the successful synthesis of Bi_2WO_6 [42]. The HRTEM image (Fig. 1j) depicts the good crystallization of 2% CSs- Bi_2WO_6 nanoparticles but an amorphous structure for CSs [43]. The above indicated that the composite catalyst was successfully prepared.

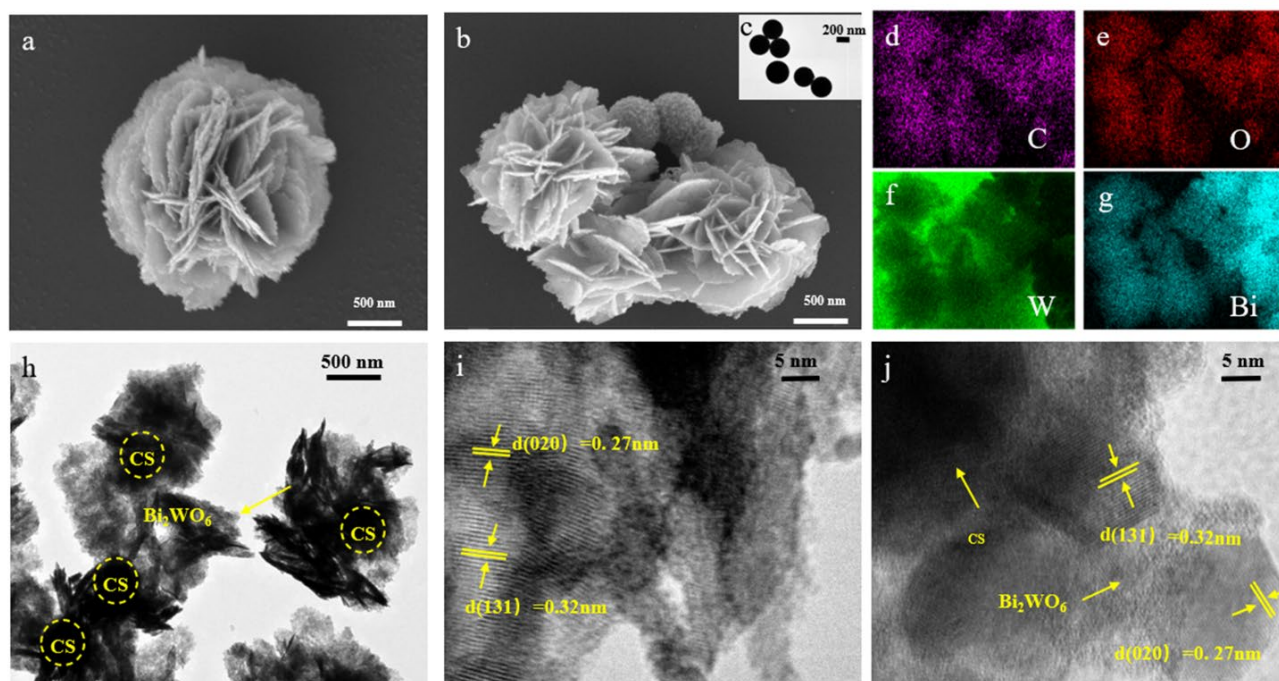


Fig. 1 SEM images of Bi_2WO_6 (a), 2% CSs- Bi_2WO_6 (b), CSs (c), the element distribution of C, O, Bi, and W in 2% CSs- Bi_2WO_6 (d-g), TEM images of 2% CSs- Bi_2WO_6 (h), and HRTEM images (i, j) of Bi_2WO_6 and 2% CSs- Bi_2WO_6

The XRD patterns showed the phase structure of the CSs- Bi_2WO_6 complexes in different proportions (Fig. 2a). The crystal plane (131), (020), (220), (313), (226), (400), (333), and (406) correspond to the characteristic diffraction peaks $2\theta = 28.14^\circ, 32.76^\circ, 47.03^\circ, 55.76^\circ, 58.44^\circ, 68.79^\circ, 75.91^\circ, \text{ and } 78.24^\circ$, respectively. These peaks are sharp and clear as well as matched well with Bi_2WO_6 standard card (PDF#26-1044), demonstrating the successful preparation of Bi_2WO_6

[44, 45]. Fig. S3 further showed the XRD image of CSs. It exhibited a broad peak at around $\sim 44^\circ$, corresponding to (101) reflections of carbon materials. In the XRD pattern of the composite, the diffraction peak of the amorphous structure of CSs phase was not found. This is to some extent expected because of the low crystallinity of amorphous carbon and the incorporation of trace carbon materials during the preparation [46].

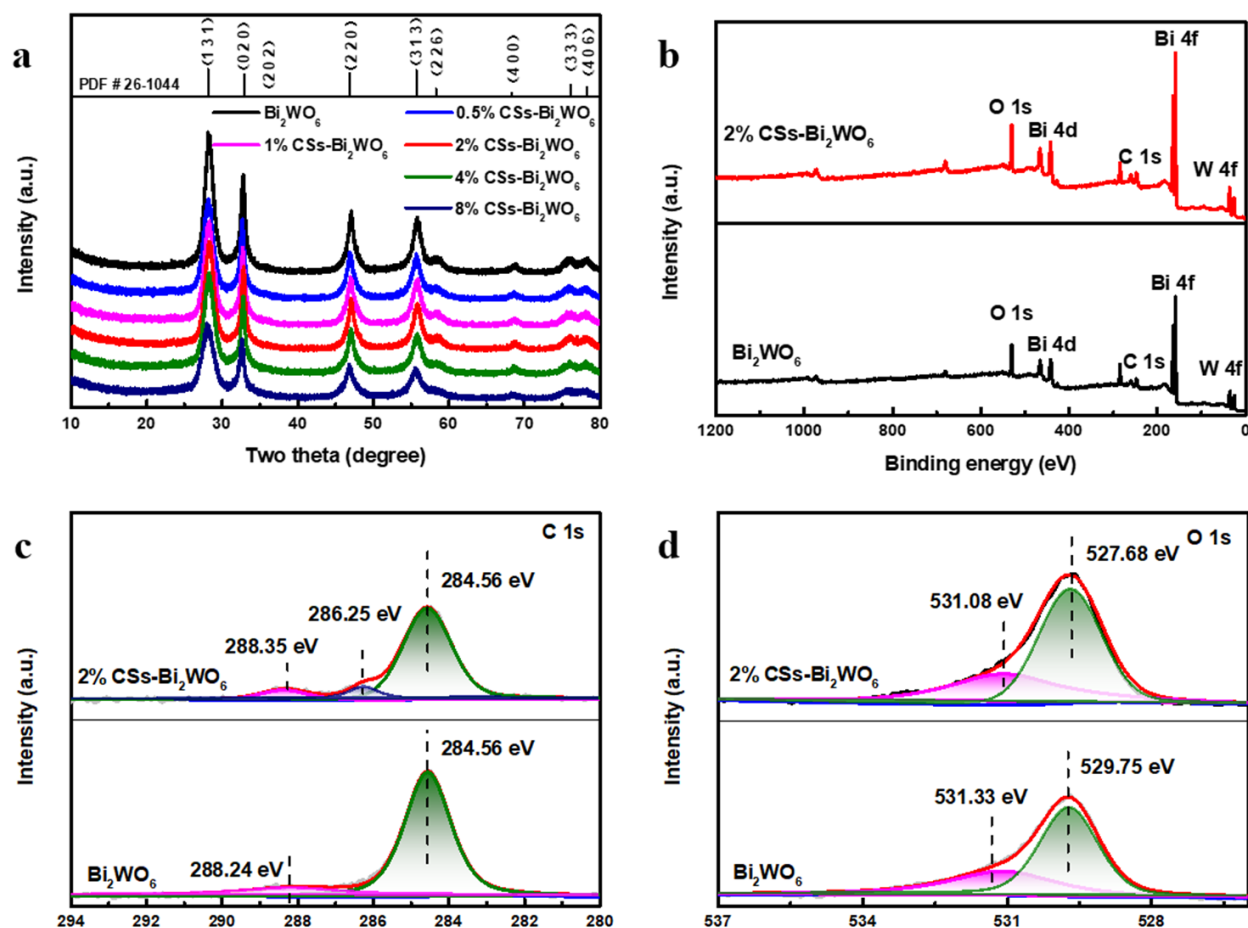


Fig. 2 XRD patterns of Bi_2WO_6 and $\text{CSs-Bi}_2\text{WO}_6$ photocatalysts (a), XPS spectra of the Bi_2WO_6 and 2% $\text{CSs-Bi}_2\text{WO}_6$ (b), C 1s (c), O 1s (d)

The chemical states of Bi_2WO_6 and 2% $\text{CSs-Bi}_2\text{WO}_6$ were analyzed by XPS. Fig. 2b shows the existence of elements C, O, Bi, and W in the sample. The elements of Bi and W peaks were found in supporting information of Fig. S4. Fig. 2c shows C 1s peak for pure Bi_2WO_6 and 2% CSs-

Bi_2WO_6 . All elements' binding energy was calibrated with C 1s of aliphatic carbon at 284.60 eV. The C 1s peak at around 288.24 eV of Bi_2WO_6 was attributed to the adsorbed CO_2 on the surface [47]. The slight shift of C 1s peak at 288.35 eV was speculated of 2% CSs- Bi_2WO_6 to be caused by C = O bond generated by the introduction of BCs. Notably, 2% CSs- Bi_2WO_6 shows a new peak at 286.2 eV, which may be due to the formation of C - O - C between Bi_2WO_6 and CSs (O atom in Bi_2WO_6). The above result indicated that most of C in 2% CSs- Bi_2WO_6 was sp^2 hybridized. The presence of oxygen-containing functional groups in the 2% CSs- Bi_2WO_6 system suggests that the strong interaction between Bi_2WO_6 and carbon nanospheres is formed during the hydrothermal reaction.

As shown in Fig. 2d, the asymmetric peak centered of O 1s at 530 eV was decomposed into two components at the binding energy of 529.74 eV and 531.33 eV for pure Bi_2WO_6 , which were due to the surface lattice oxygen and the adsorbed oxygen species respectively [48, 49]. With the addition of the CSs, the binding energy shifts from 529.75 eV, 531.33 eV (Bi_2WO_6) to 531.08 eV, 527.68 eV (2% CSs- Bi_2WO_6). The peaks of O 1s of 2% CSs- Bi_2WO_6 shifted to the lower binding energy compared to Bi_2WO_6 , indicating the electron cloud density and electronegativity around O decrease due to the addition of carbon spheres. the interactions between Bi_2WO_6 and BCs in the composite [50, 51].

Nitrogen sorption measurements evaluated the surface area and pore structural properties in Fig. 3. In Fig. 3a, the isotherm curves belonged to type IV with H3 hysteresis loops [43, 52, 53]. And specific surface area of pure Bi_2WO_6 , 0.5% CSs- Bi_2WO_6 , 1% CSs- Bi_2WO_6 , 2% CSs- Bi_2WO_6 , 4% CSs- Bi_2WO_6 , and 8% CSs- Bi_2WO_6 was 20.76, 82.45, 84.80, 84.88 85.04 and 82.31 $\text{m}^2 \text{g}^{-1}$,

respectively. Obviously, the addition of surface area was mainly caused by the surface folds of CSs, which will increase the contact area of CSs with contaminants. Interestingly, the surface area of the complex actually decreased when an excess of CSs was introduced. This may be due to the fact that with the increase of CSs, the packing of Bi_2WO_6 became dense, the specific surface area decreased slightly, and the active site decreased [54, 55]. This proves that adsorption is strongly related to the specific surface area of the composite catalyst. The pore size distribution is shown in Fig. 3b, the pore size of samples is almost distributed from 1 nm - 20 nm. Furthermore, the slit-shaped pore structure displayed by Bi_2WO_6 is consistent with the sheet morphology shown by SEM results.

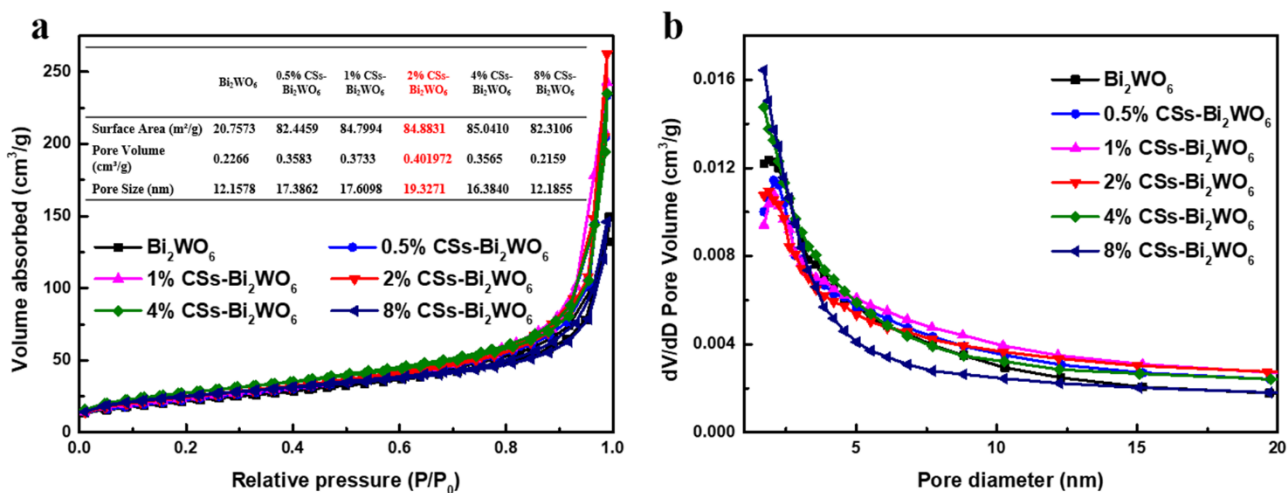


Fig. 3 N_2 adsorption-desorption isotherms (a), pore size distribution curves (the inset) of Bi_2WO_6 and CSs- Bi_2WO_6 (b)

3.2 Optical and electronic properties

It is well known that the utilization of visible light mainly depends on the band gap (E_g) and energy level of the semiconductor photocatalyst. The optical absorption abilities and bandgap of prepared samples were measured by DRS in Fig. 4.

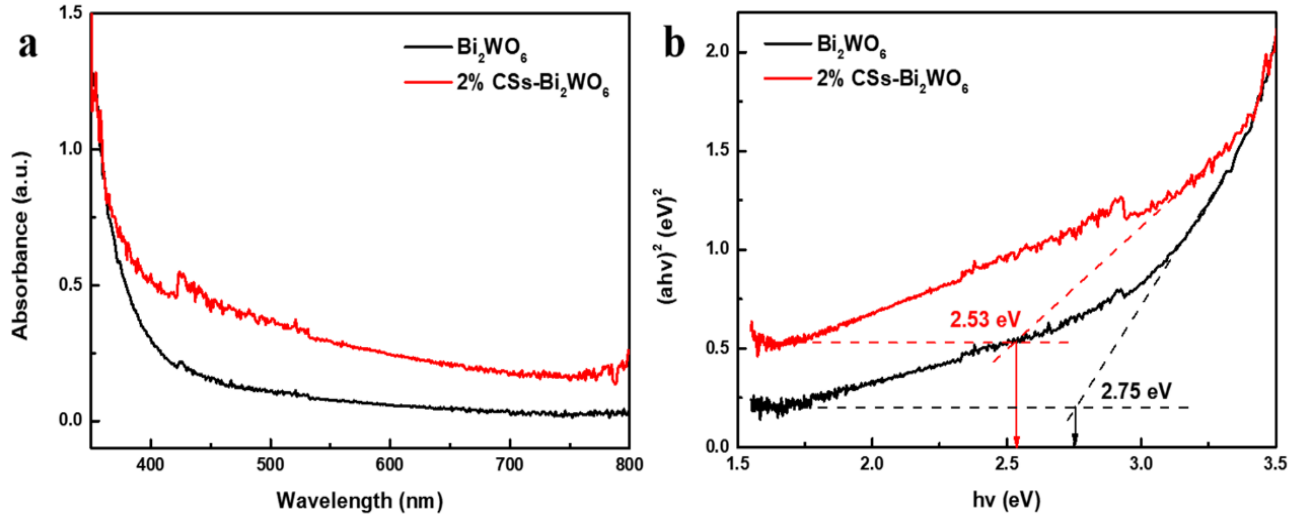


Fig. 4 UV-vis diffuse reflectance spectra of Bi_2WO_6 and 2% CSs- Bi_2WO_6 (a), plot of Kubelka - Munk transformation of Bi_2WO_6 and 2% CSs- Bi_2WO_6 (b)

Fig. 4a shows the comparison of UV-VIS DRS optical absorption characteristics between pure Bi_2WO_6 and 2% CSs- Bi_2WO_6 . It can be observed that the optical absorption capacity of 2% CSs- Bi_2WO_6 at 300-800 nm wavelength is higher than that of Bi_2WO_6 . In addition, after the introduction of CSs, the light absorption capacity of the sample has undergone a significant redshift. It shows that the composite photocatalyst has higher light utilization efficiency, promotes the transfer of charge, and improves photocatalytic performance. The band gaps energies (E_g) were calculated as the following formula [41]:

$$ahv = A(hv - E_g)^{\frac{n}{2}} \quad (1)$$

where A represents a constant, and α , ν , and E_g represent the adsorption coefficient, optical frequency, and bandgap, respectively [56]. The electronic transition type of the semiconductor determines the value of n [57]. Referring to previous studies, Bi₂WO₆ is an indirect bandgap semiconductor with an N value of 4 [58].

In Fig. 4b, the band gap of the pure Bi₂WO₆ was estimated to be 2.75 eV, which was consistent with the reported experimental values [59]. The band gap of the 2% CSs-Bi₂WO₆ sample was about 2.53 eV, which was lower than that of Bi₂WO₆.

Photoluminescence (PL) spectra is considered to be an important experiment to discuss the efficiency of e⁻ and h⁺ pair separation. Fig. 5a showed the strong emission peaks of the samples at 425 nm under excitation at 378 nm. In general, the emission intensity of CSs-Bi₂WO₆ was lower than that of pure Bi₂WO₆, and the 2% CSs-Bi₂WO₆ showed the lowest emission intensity, which proved its better photocatalytic activity. This trend suggested that the separation efficiency of photogenerated e⁻ and h⁺ can be effectively improved by introducing CSs, which promotes molecular diffusion and transfer.

In Fig. 5b, the flat-band (C_{FB}) potential of the composite photocatalyst is derived from the X-intercept of the linear region. And the obtained results can be further transformed into NHE potential [29, 60, 61]. The straight upward curves demonstrated that Bi₂WO₆ was an n-type semiconductor. The C_{FB} potentials of both the Bi₂WO₆ and 2% CSs-Bi₂WO₆ composites was -0.55 V vs. Ag/AgCl (-0.35 eV vs. NHE). The E_{CB} for both the Bi₂WO₆ and 2% CSs-Bi₂WO₆ Schottky contacts was approximately -0.45 eV. Combined with the E_g values obtained from Tauc plots in Fig. 4b, the valence band potentials (VB) of the 2% CSs-Bi₂WO₆ sample was calculated

to be 2.08 eV [62].

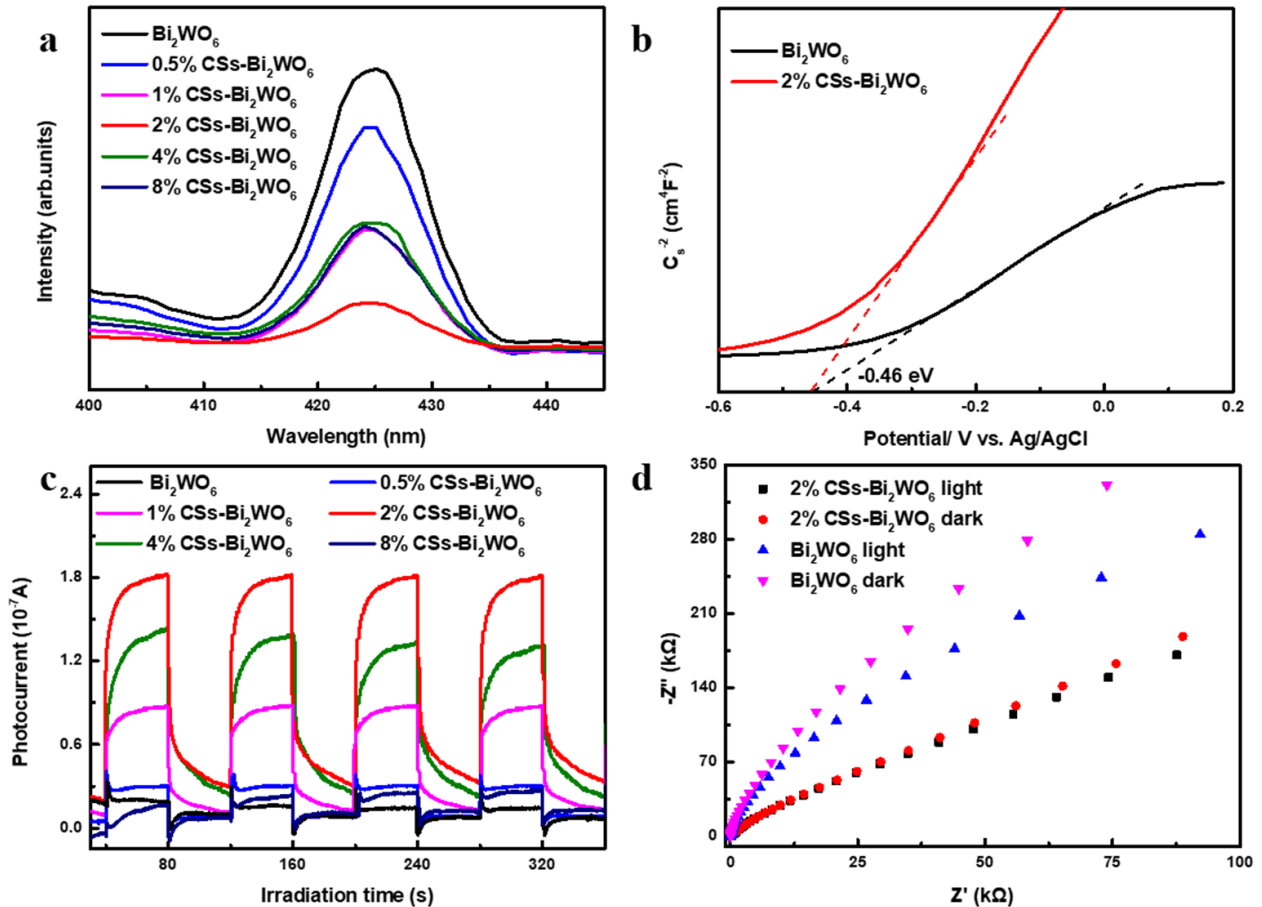


Fig. 5 PL spectra (a), Mott-Schottky curve of Bi_2WO_6 and 2% $\text{CSs-Bi}_2\text{WO}_6$ (b), transient photocurrent response of $\text{CSs-Bi}_2\text{WO}_6$ (c), electrochemical impedance spectroscopy (EIS) measurements of Bi_2WO_6 and 2% $\text{CSs-Bi}_2\text{WO}_6$ (d)

The photocurrent responses of the pure Bi_2WO_6 and 2% $\text{CSs-Bi}_2\text{WO}_6$ were tested and the results are displayed in Fig. 5c. Obviously, both the photoelectrodes of all samples show stable photocurrent responses over several switching cycles. All samples immediately produce a photocurrent when the electrode is momentarily exposed to visible light. On account of the higher recombination rate of photogenerated e^- and h^+ in the Bi_2WO_6 crystal, it can be found that Bi_2WO_6 had lower photocurrent responses. The photocurrent response of the composite photocatalyst 2% $\text{CSs-Bi}_2\text{WO}_6$ was about 9 times higher than pure Bi_2WO_6 , which effectively indicated more

efficient photo-induced charge separation and faster electron transport. Fig. 5d shows the EIS Nyquist plots for Bi_2WO_6 and 2% CSs- Bi_2WO_6 under different light conditions, respectively. The graph demonstrated that the arc radius of Bi_2WO_6 was larger than 2% CSs- Bi_2WO_6 , indicating that the introduction of CSs enhanced the charge migration of Bi_2WO_6 and reduces the reaction resistance at the semiconductor interface. Furthermore, the EIS semicircle radius of 2% CSs- Bi_2WO_6 under light is smaller than its under dark condition, which indicates that light can excite more charge carriers and enhance photocatalytic activity [63].

3.3 Photocatalytic tests

The photodegradation result of samples are shown in Fig. 6. TC was selected as the target pollutant to test its photocatalytic performance. Under visible light, 50 mg photocatalyst was added to 100 mL 50 mg / L TC solution for photocatalytic degradation. Fig. 6a is a diagram of the removal rate of TC by the catalyst. The composite photocatalyst reached adsorption-desorption equilibrium within 40 minutes of the dark reaction (adsorption kinetics data in Fig. S5). The degradation rate of TC by 2% CSs-Bi₂WO₆ was 84.6%, which is nearly 25% higher than pure Bi₂WO₆. This was because 2% CSs-Bi₂WO₆ had a large specific surface area, especially the folded structure of CSs, which effectively increased the adsorption of TC by the catalyst and promoted photoreaction. Notably, we found that the degradation effect of 0.5% CSs-Bi₂WO₆ was slightly higher than that of 1% CSs-Bi₂WO₆. In order to further explore its regularity, we added the photocatalytic degradation data of TC by 0.2% CSs-Bi₂WO₆ and 0.8% CSs-Bi₂WO₆ in the supporting information (Fig. S6) [64, 65]. However, even with the addition of a trace amount of CSs, the photocatalytic performance of the composite photocatalyst is still higher than that of pure BW, which indicates that CSs can effectively enhance the utilization of visible light and form a strong interfacial electronic effect with semiconductor and then improve the photocatalytic efficiency. Compared with the work reported in the literature on the degradation of organic pollutants by carbon materials in Table S2.

Fig. 6b shows the quasi-first-order degradation kinetics of the catalyst composite with a linear relationship between irradiation time and $\ln(C_0/C_t)$. The degradation rate constant was worked

from the slope of the kinetic curve. The rate constants for Bi_2WO_6 , 0.5% CSs- Bi_2WO_6 , 1% CSs- Bi_2WO_6 , 2% CSs- Bi_2WO_6 , 4% CSs- Bi_2WO_6 , and the 8% CSs- Bi_2WO_6 were 0.01502 min^{-1} , 0.00257 min^{-1} , 0.02141 min^{-1} , 0.03929 min^{-1} , 0.02951 min^{-1} and 0.0703 min^{-1} , respectively. In particular, 2% CSs- Bi_2WO_6 had the highest photocatalytic degradation rate, which should be put down to more efficient charge separation by adding CSs. The stability and reproducibility of photocatalysts are of great research value in practical applications. Fig. 6c shows the photocatalytic cycling test plots of the prepared catalyst samples. After the 5th cycle, the final degradation rate was about 80.5%, which displayed better photocatalytic stability of 2% CSs- Bi_2WO_6 .

We used different capture scavengers under visible light to explore the main active substances. In the present study: dimethyl sulfoxide (DMSO), ammonium oxalate (AMO), 1,4-benzoquinone (BQ), and isopropyl alcohol (IPA) were used as electron (e^-), hole (h^+), superoxide radical ($\cdot\text{O}_2^-$), and hydroxyl radical ($\cdot\text{OH}^-$) scavengers, respectively [66]. In Fig. 6d. the presence of BQ and AMO significantly inhibited the photodegradation of TC, indicating that $\cdot\text{O}_2^-$ was the main active species. Photodegradation was also somewhat inhibited when AMO was added into the system, indicating that the h^+ was more active in the composite photocatalyst. In addition, the addition of IPA and DMSO had little effect on the photocatalytic performance. In short, the order of action of the active species in the 2% CSs- Bi_2WO_6 photocatalytic degradation TC was $\cdot\text{O}_2^- > h^+ > e^- > \cdot\text{OH}^-$

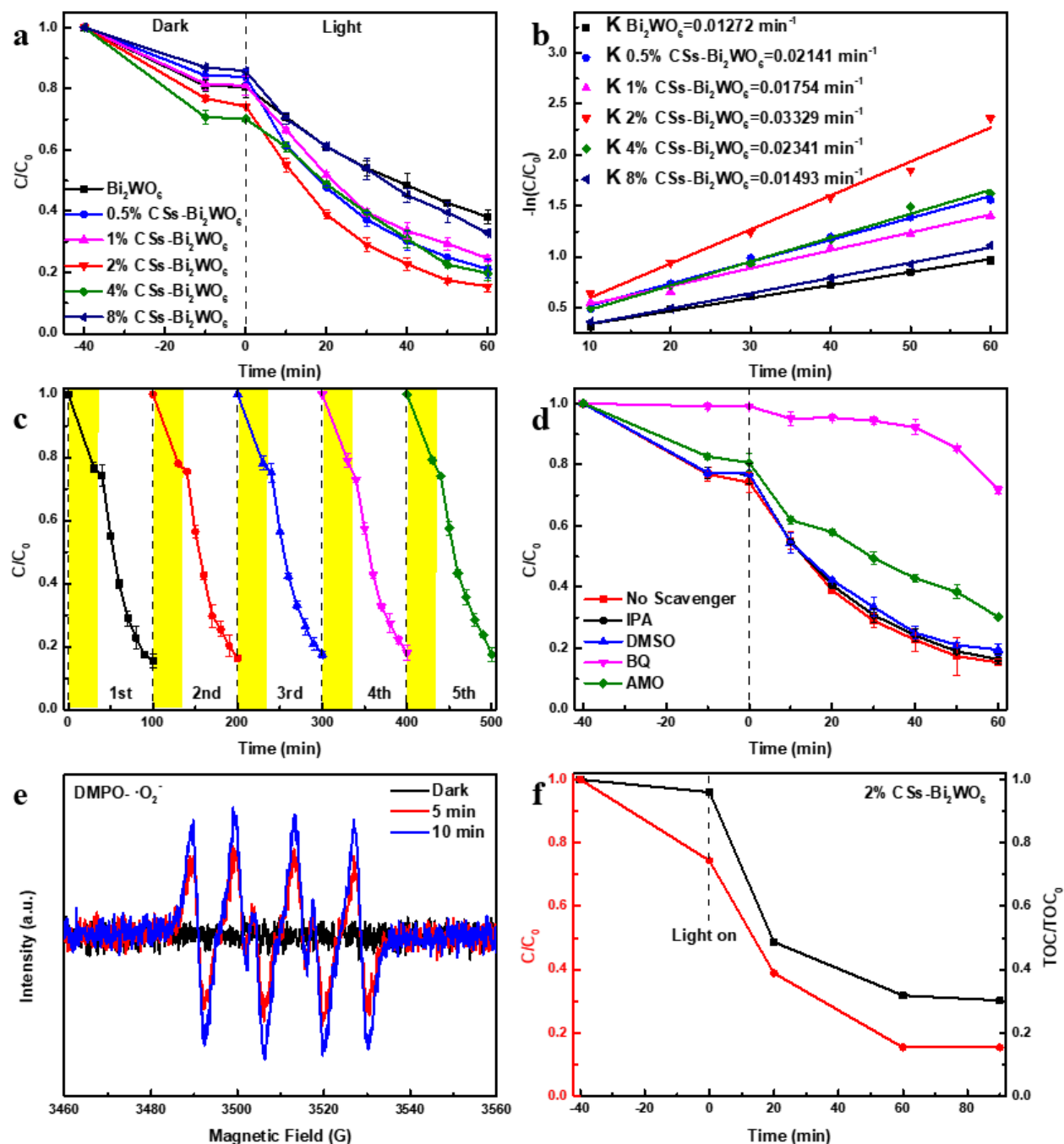


Fig. 6 Variation of TC concentration with irradiation time under light irradiation. (The experiment was repeated three times) (a), first-order kinetic curves of photocatalytic TC degradation in the as-synthesized samples (b), reusability of 2% CSs- Bi_2WO_6 in five runs for TC (c), mapping of active species captured in the TC photodegradation system on 2% CSs- Bi_2WO_6 under visible light (d), ESR spectrum of DMPO- $\cdot\text{O}_2^-$ (e), removal of TOC as a function of time for TC degradation using as-prepared 2% CSs- Bi_2WO_6 (f)

In Fig. 6e, a further identification test of the free radical O_2^- was carried out using electron

spin resonance (ESR). No ESR signal was observed in the dark condition. When 2% CSs-Bi₂WO₆ and DMPO were mixed under light for 5 and 10 minutes, a set of characteristic peaks attributed to ·O₂⁻ was found. This indicates that O₂⁻ was produced under light conditions in the catalytic system. The signal intensity of ·O₂⁻ increased with the increase of irradiation time. Relevant studies have shown that 2% CSs-Bi₂WO₆ photocatalytic TC not only has a higher degradation rate, but also a higher degree of organic matter mineralization [9, 67, 68]. Therefore, we studied the removal efficiency of total organic carbon (TOC), and the results are shown in Fig. 6f. It seems that in the 90 min degradation process, the reduction curve of TOC is mainly based on the degradation curve, and the degradation rate is 68.17%. TOC ratio was slightly higher than residual antibiotic concentration ratio, indicating that some intermediates were not completely degraded or degraded for a long time.

3.4 Photocatalytic mechanism

The intermediates of 2% CSs-Bi₂WO₆ photodegradation of TC were identified by LC-MS technology, and the degradation pathway of TC was further described. 12 intermediates with m/z values of 481, 475, 459, 445, 415, 388, 344, 318, 300, 274, 250, 246, 242, and 165 were generated under visible light irradiation. According to Fig. 6d, ·O₂⁻ was the main active substance in the degradation process. Therefore, the photodegradation pathway of TC can be proposed (Fig. 7 and Fig. S7) [69, 70].

After visible light irradiation, TC was invaded from three possible sites. Pathway 1: The main reaction pathway of TC degradation is the reaction of superoxide radical attacking phenol group. Under light conditions, it was further oxidized to form a macromolecular intermediate with m/z = 475. Then, the product with m/z = 475 was converted to the product with m/z = 459 by attacking the ketone group and changing the hydroxyl group [71]. Finally, the unstable intermediate cleaved due to the loss of the hydroxyl group, resulting in the formation of small molecules with m/z of 344, 300 and 242, respectively. They are eventually mineralized into H₂O and CO₂. Pathway 2: The formation of m/z = 481 is due to the attack of the double bond of TC, which further introduces hydroxyl and ketone groups [72]. In the next period, the product of m/z = 246 and 165 appears due to the opening of the benzene ring by h⁺ attack. Eventually mineralized into small molecules [73]. Pathway 3: the product of m/z = 338 was attributed to the loss of the N-methyl group. The product with m/z = 318 is formed due to further cleavage of the C-ring and loss of the hydroxyl group. Furthermore, the small

molecule with $m/z = 318$ was transformed into a product with $m/z = 274$ by removing the carboxyl group. Finally, after different process of detachment of hydroxyl, and aldehyde groups followed by decarbonization and dihydroxylation [74]. With increasing reaction time, oxidative decomposition and ring-opening reactions further occur, and the above intermediate products are finally decomposed into H_2O , CO_2 and NH_4^+ , etc.

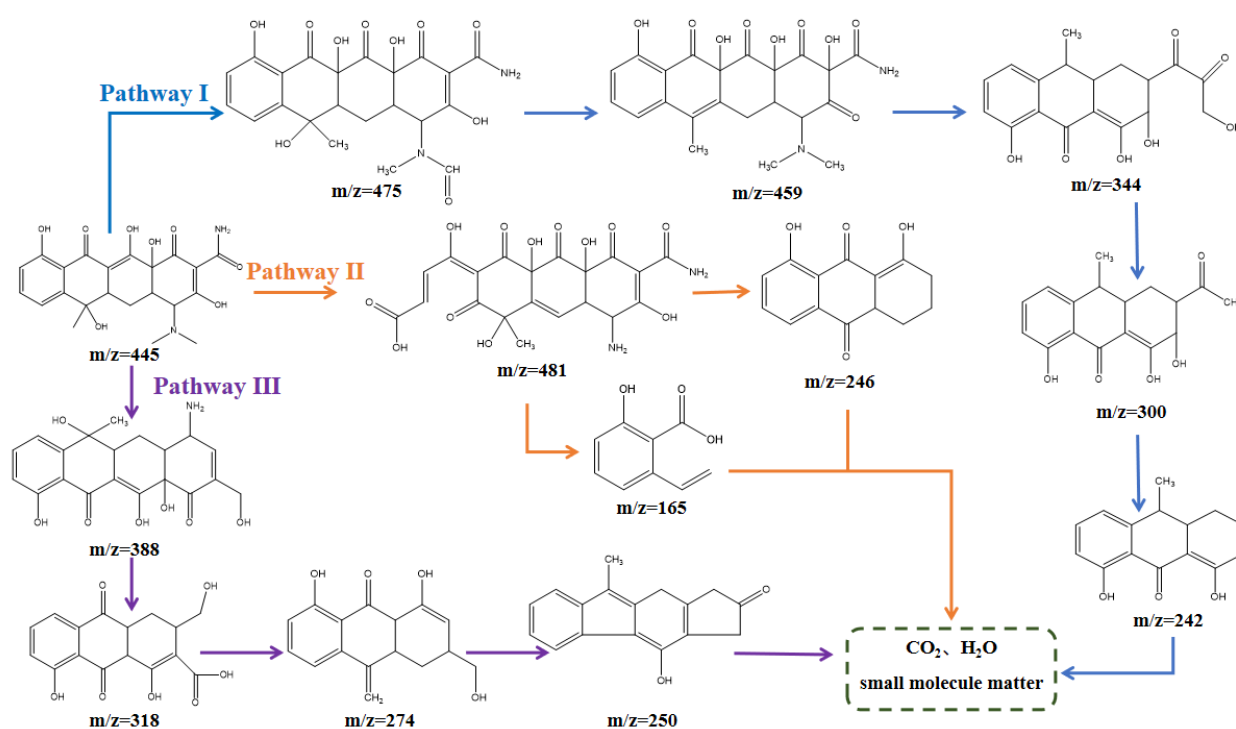
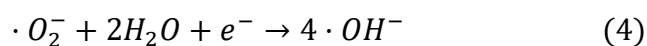
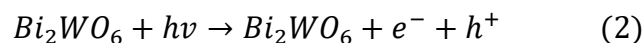


Fig. 7 Degradation pathway of TC over 2% CSs-Bi₂WO₆

According to the above experiments, Fig. 8 shows the mechanism of photocatalytic degradation of TC by 2% CSs-Bi₂WO₆ under visible light irradiation. The CB of 2% CSs-Bi₂WO₆ was - 0.45 eV, which was more negative than the redox potential ($E_{O_2/O_2^-} = - 0.33$ eV). Thus, the main reason for TC degradation was that the photogenerated electrons were transferred to O₂ and further converted to $\cdot O_2^-$. In

addition, combined with the E_g energy of 2% CSs-Bi₂WO₆, the VB potential of 2% CSs-Bi₂WO₆ can be deduced as + 2.08 eV according to the equation ($E_{CB} = E_{VB} - E_g$) [75]. Specific data are described in Fig. 4. The VB level of 2% CSs-Bi₂WO₆ was lower than $E_{\cdot OH/OH^-} = + 2.31$ eV. Therefore, $\cdot OH^-$ did not play a major role in the photocatalytic reaction. Under visible light irradiation, the semiconductor photocatalyst generates electron-hole pairs. The photogenerated electrons reacted with O₂ (from the air) on the photocatalyst surface to form the $\cdot O_2^-$. TC is eventually mineralized into small molecule compounds, such as water and carbon dioxide [38, 76]. To sum up, the 2% CSs-Bi₂WO₆ photocatalyst has higher electron-hole separation efficiency than the bare Bi₂WO₆ photocatalyst. Further, the photodegradation reaction process of TC has been proposed as Equations (2-6):



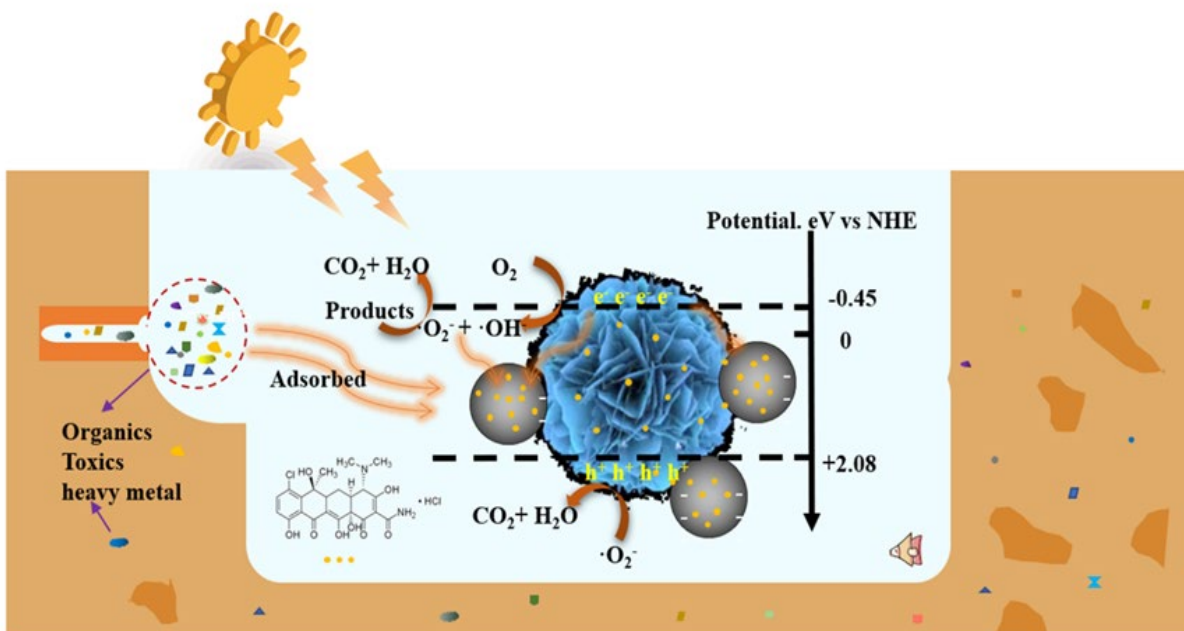


Fig. 8 Possible mechanism of TC on the surface of 2% CSs-Bi₂WO₆ composite

During the dark reaction, TC was adsorbed on the surface of CSs due to irregular folds on the surface and strong adsorption. It facilitated the transfer of pollutants from the environment to the surface of the catalyst, forming a high concentration TC environment on the surface. And finally adsorption - dissolution equilibrium was achieved. In addition, CSs are more efficient at harvesting light by multiple scattering and utilizing photons. The composite can effectively separate the e^- / h^+ pairs when photogenerated electrons in bismuth tungstate migrate to the CSs, thus improving the photocatalytic performance.

4 Conclusion

In summary, the composite photocatalyst of CSs-Bi₂WO₆ was synthesized by a simple hydrothermal method. The experimental results showed that the composite showed good photocatalytic performance (84.6%) and stability for tetracycline degradation under the optimal CSs loading amount (2 wt%). The degradation kinetics of TC in CSs modified Bi₂WO₆ was 2.6 times that of pure Bi₂WO₆. This may be because CSs can enhance the absorption of visible light by the composite catalyst and effectively enhance the separation efficiency of e⁻/h⁺. It is worth noting that the TOC estimation of the post-degradation reaction medium corresponded to 68.2% mineralization. Based on the capture experiment and ESR measurement results, the free radical ·O₂⁻ plays a leading role in the photocatalytic reaction. Further, the photocatalytic reaction mechanism and degradation pathway of 2% CSs-Bi₂WO₆ were revealed in detail by using LC-MS determined reaction intermediates. This study provides a new idea for photocatalytic degradation of organic pollutants through the synergistic action of carbon and semiconductor materials.

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Author Contributions Statement

Zhanhua Huang: conceptualization, methodology, writing - review and editing, Funding acquisition. Xiaona Jiang, Xinrui Zhang, and Houjuan Qi wrote the main manuscript text. Shuai Chen and Lanni Qu prepared figures. All authors reviewed the manuscript.

Declaration of Competing Interest

There are no conflicts to declare.

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